

**Politecnico Di Milano**  
Environmental and Geomatic Engineering



**“Source Apportionment of PM<sub>2.5</sub> in  
Milan by PMF Receptor Model”**

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# Table of Contents

Table of Contents.....	3
List of Figures .....	4
List of Tables .....	5
Abstract.....	6
1 Introduction .....	7
1.1 Air Pollution .....	8
1.2 Air Pollution Occurrences .....	9
1.3 Type of Air Pollution .....	11
1.4 Particulate Matter.....	13
1.5 Trends and Projections .....	18
1.6 Need for Source Apportionment .....	22
1.7 Scope of the Work .....	27
2 Model Application and Material .....	28
2.1 PMF Application.....	28
2.2 PMF within Europe.....	30
2.3 PMF Model.....	32
2.3.1. Comparison to PMF 3.0.....	33
2.4 Input data.....	34
2.5 Output Data .....	35
2.6 Environmental Data .....	37
3 Results and Discussion.....	41
3.1 PMF Objective.....	41
3.3.1 PMF Application.....	41
3.3.2 PMF Profiles .....	43
3.3.3 Comparison to Other Receptor Models.....	56
Conclusion.....	62
Literature .....	67
Annex I – How PMF Works.....	72

## List of Figures

- Figure 1: Global Wind Trends
- Figure 2: Temperature Inversion
- Figure 3: Primary Pollutants Share
- Figure 4: Particulate Matters Size
- Figure 5: Global Trend of PM<sub>10</sub> and PM<sub>2.5</sub>
- Figure 6: Global Satellite Deriver Map of PM<sub>2.5</sub>, averaged over 2001-2006
- Figure 7: Yearly Total Yield of Estimated Solar Electricity Generation (kWh)
- Figure 8: Yearly Total Yield of Estimated Production Potential of Wind Power Stations
- Figure 9: Schematic Representation of the Different Methods for Source Identification
- Figure 10: Percentage of Model Types Used for Source Apportionment by Different EU Countries
- Figure 11: Time Trend of RM Studies in Europe between 2001 and 2010
- Figure 12: Sampling Site
- Figure 13: Reconstructed PM<sub>2.5</sub> Concentrations vs. Measured PM<sub>2.5</sub>
- Figure 14: Reconstructed PM<sub>2.5</sub> Concentrations vs. Measured PM<sub>2.5</sub> - Warm and Cold Season, respectively
- Figure 15: Concentration Time Series
- Figure 16: Factor Contribution
- Figure 17: Factor Fingerprint
- Figure 18: Source Concentration (temporal series) ( $\mu\text{g}/\text{m}^3$ )
- Figure 19: Factor Contribution Warm Season
- Figure 20: Fingerprint Warm Season
- Figure 21: Source Concentration Warm Season (temporal series) ( $\mu\text{g}/\text{m}^3$ )
- Figure 22: Factor Contribution Cold Season
- Figure 23: Fingerprint Cold Season
- Figure 24: Source Concentration Cold Season (temporal series) ( $\mu\text{g}/\text{m}^3$ )
- Figure 25: Box plot for the PM<sub>2.5</sub> Mass Contribution
- Figure 26: Daily to Mean Ratios for all Sources in the Annual Dataset
- Figure 27: Summary of UNMIX and PMF Sources, and Average Contribution to PM<sub>2.5</sub> for Annual Dataset ( $\mu\text{g}/\text{m}^3$ )
- Figure 28: Summary of UNMIX, PMF and CMB Sources, and Average Contribution to PM<sub>2.5</sub> for Warm Season Dataset ( $\mu\text{g}/\text{m}^3$ )
- Figure 29: Summary of UNMIX, PMF and CMB Sources, and Average Contribution to PM<sub>2.5</sub> for Cold Season Dataset ( $\mu\text{g}/\text{m}^3$ )
- Figure 30: Conjugate Gradient Method

## List of Tables

- Table 1: Atmospheric Chemical Composition
- Table 2: Contribution by Renewable Energy Carriers (electricity, heating/cooling, transport)
- Table 3: Input Concentration File
- Table 4: Dataset, Via Messina, Milan
- Table 5: Dataset after Reduction
- Table 6: Principal Statistical Parameters of Samples
- Table 7: Factor Profiles ( $\mu\text{g}/\text{m}^3$ )
- Table 8: Factor Profiles Warm Season ( $\mu\text{g}/\text{m}^3$ )
- Table 9: Factor Profiles Cold Season ( $\mu\text{g}/\text{m}^3$ )

## Abstract

This case study discusses the application of a multivariate receptor model, the EPA PMF 5.0 to the PM<sub>2.5</sub> dataset from Lombardy region in Italy. The aim of the study is to perform source apportionment investigation of the applied dataset and identify different PM<sub>2.5</sub> sources that greatly impact the composition of particulate matter in the studied region.

PMF model evaluates contribution to diverse source types of measured PM<sub>2.5</sub> concentrations by investigating chemical composition of ambient pollution samples. As a type of receptor models, PMF used as an input data, PM concentrations and their relative chemical specification and provides as an outcome the number of sources, their composition and the source contributions.

The analysis has been performed to dataset which is comprised of PM<sub>2.5</sub> sampling campaign performed in the downtown Milan between 2002 and 2003. The original data set is consisting of 162 daily samples of PM<sub>2.5</sub> mass concentration and relative chemical specification of 21 chemical species (carbon components, inorganic ions and trace elements). However, as some samples did not contain measurements for all species, and this represent the main requirement for model to be run, the original dataset had to be reduced. Likewise, reduced dataset consisted of 99 daily samples of PM<sub>2.5</sub> mass concentration and 11 chemical species.

The analysis of total annual PM<sub>2.5</sub> mass concentration revealed presence of 6 sources (secondary sulfate, traffic non-exhaust, biomass combustion/break wear, domestic heating, re-suspended soil dust and secondary nitrate). After the general examination, the dataset was split into two subsets, warm and cold season for the more detailed study. The warm season analyses identified 4 sources (secondary nitrates and organics, biomass combustion/break wear, traffic exhaust and secondary sulfate), while on the other hand the cold season identified 4 sources (secondary nitrates and organics, domestic heating, crustal matter and re-suspended soil dust).

## **1 Introduction**

Human being can endure hours without the water, days without the food, but only few minutes without the air. We must have air to live. However, breathing polluted air can cause severe health problems and in some cases even death.

On the other hand, contamination of the air is also damaging natural environment. Trees, crops, rivers, lakes and animals are intensely influenced. Its sustainability and diversity is rapidly changing, accompanying the rate of change of world's pollution.

The pollution is present in many forms and it represents threat to human being in this modern world. The water we drink the air we breathe, the ground where we grow out food, and even the noise we hear every day. All these elements contribute to severe health problems and a lower quality of life. Hence, our awareness of the problem should be increased. Humans should not only understand the air pollution but also they should know how to manage air quality. People should know how deeply they are affected in their daily life routines. What the methods are in order to decrease and prevent its presence, and how they can protect themselves from serious consequences. Additionally, the sources of air pollution must be identified, their effects to our ecosystem, mechanisms to its reduction and monitoring technologies on a local and global scale.

## 1.1 Air Pollution

The present-day atmosphere is entirely different from the natural atmosphere that existed before the Industrial Revolution (18<sup>th</sup> century) in terms of chemical composition. If the natural atmosphere is deliberated as “clean”, than this means that in today’s atmosphere clean air cannot be found anywhere. In the Table 1 the chemical composition of the pre-industrial natural atmosphere is compared to current composition of the atmosphere.

Table 1: Atmospheric Chemical Composition (CDIAC, May 2014)

Gas	Pre-1750 tropospheric conc.	Recent tropospheric conc.	GWP(100 y. time horizon)	Atmospheric lifetime (years)	Increased radiative forcing (W/m <sup>2</sup> )
<b>Concentration in parts per million (ppm)</b>					
<b>Carbon dioxide (CO<sub>2</sub>)</b>	280	401.33	1	100-300	1.88
<b>Concentration in parts per billion (ppb)</b>					
<b>Methane(CH<sub>4</sub>)</b>	722	1893	28	12	0.49
<b>Nitrous oxide (N<sub>2</sub>O)</b>	270	326	265	121	0.17
<b>Tropospheric ozone(O<sub>3</sub>)</b>	237	337	n.a.	Hours-days	0.4

Giving a definition to “Air pollution” is not simple. There is a common opinion that air pollution started when humans began burning fuels. Particularly, all man-made (anthropogenic) emissions may be called air pollution, since they alter the chemical composition of the natural atmosphere. The increase in the global concentration of greenhouse gases CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and H<sub>2</sub>O (shown in Table 1) can be called air pollution even though is not found that their concentration is toxic for humans and the ecosystem. Redefined approach considers anthropogenic emissions of harmful chemicals as air pollution (Gordon, et al. 2014).

However, there is a question, what “harmful” stands for? Harmful can mean an adverse effect on the health of living beings, an adverse effect on natural or anthropogenic non-living structures. Likewise, a chemical that does not cause any short-term harmful effects may accumulate in the atmosphere during time and develop a long-term harmful effect.



The modern definition for the air pollution states that it represents any substance emitted into the air from an anthropogenic, biogenic or geogenic source<sup>1</sup>, that is present in the higher concentrations than the natural atmosphere, and may cause short-term or long-term adverse effects (Zaneti et al., 2007).

## 1.2 Air Pollution Occurrences

Air pollution represents a major health risk factor across the globe. Its complexity reflects in a fact that it cannot be easily controlled, since a lot of factors are driving it.

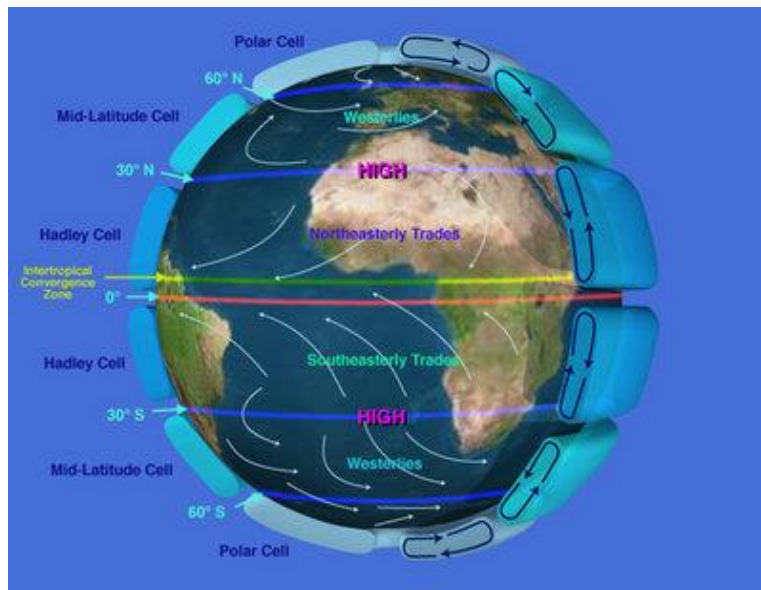
The most obvious factor influencing air pollution is the quantity of contaminants emitted into the atmosphere. However, trends in air pollution are not caused by a drastic increase in the output of pollutants; instead these trends are driven by changes in certain atmospheric conditions.

Two of the most important atmospheric conditions affecting the dispersion of pollutants are:

1. the strength of the wind and
2. the stability of the air

The direct effect of wind speed is to influence the concentration of pollutants. Referring to that, global wind trends are represented on the Image 1.

**Figure 1: Global Wind Trends (Global sailing weather, May 2014)**



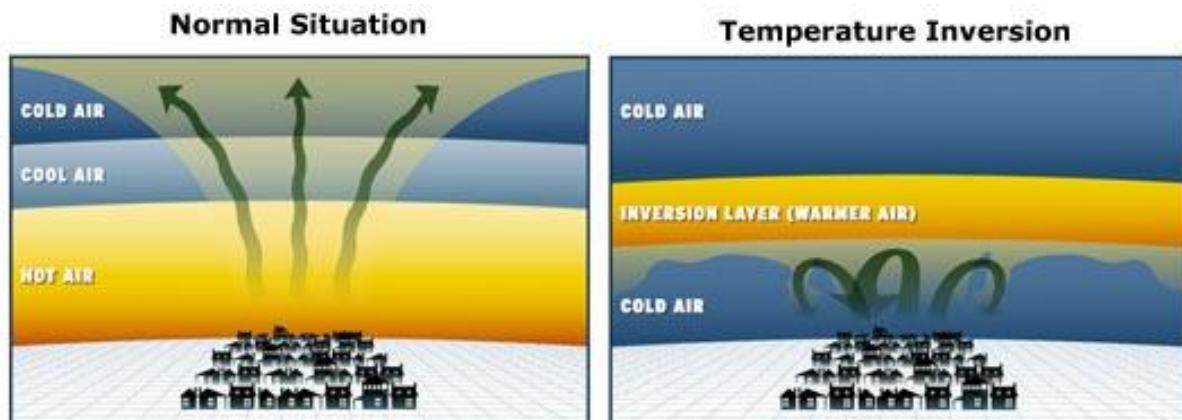
<sup>1</sup> Geogenic emissions are produced by non-living world, e.g. natural fires, volcanic emission etc.

On the contrary, atmospheric stability determines the extent to which vertical motions will mix the pollution with cleaner air above the surface layers (Pateraki et al. 2014). The vertical distance between Earth's surface and the height to which convective movements extend is called Mixing depth. Generally, the greater the mixing depth, the better the air quality is (Csavina et al. 2014).

Another important factor that influences the distribution of the pollutants is temperature inversion. During most days, the temperature of air in the atmosphere is cooler the higher up in altitude you go. The warm air rises in the atmosphere, where it expands and cools. Sometimes, however, the temperature of air actually increases with height. The situation of having warm air on top of cooler air is referred to as a temperature inversion, because the temperature profile of the atmosphere is inverted from its usual state (Wanning et al. 2014).

The most common aspect in which surface inversions form is through the cooling of the air near the ground at night. Once the sun goes down, the ground loses heat very quickly, and this cools the air that is in contact with the ground. However, since air is a very poor conductor of heat, the air just above the surface remains warm. Conditions that favor the development of a strong surface inversion are calm winds, clear skies, and long nights. Calm winds prevent warmer air above the surface from mixing down to the ground, and clear skies increase the rate of cooling at the Earth's surface. Since the nights in the wintertime are much longer than nights during the summertime, surface inversions are stronger and more common during the winter months (Tang and Al-Ajmi, 1997).

Figure 2: Temperature Inversion (Science unraveled, May 2014)



Surface temperature inversions play a major role in air quality. The warm air above cooler air acts like a lid, suppressing vertical mixing and trapping the cooler air at the surface. As pollutants from vehicles, fireplaces, and industry are emitted into the air, the inversion traps these pollutants near the ground, leading to poor air quality. The strength

and duration of the inversion will control air quality impact levels near the ground. Strong inversion will confine pollutants to a shallow vertical layer, leading to high impacts on the air quality, while a weak inversion will lead to lower impact levels. (NOAA, May 2014).

### 1.3 Type of Air Pollution

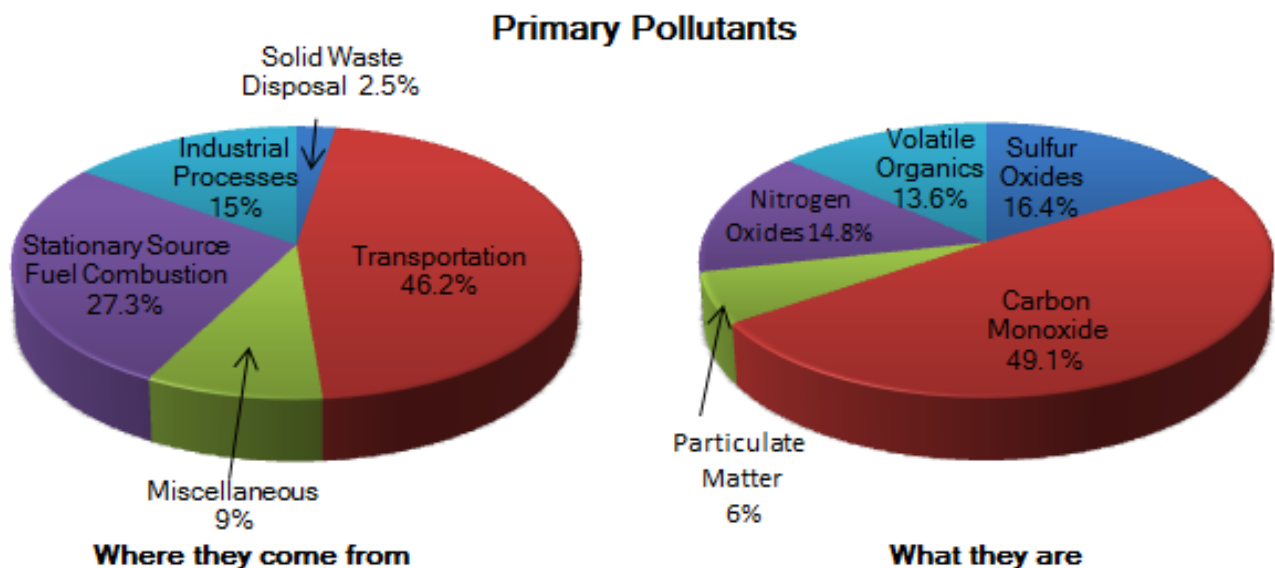
Air pollutants are any gas, liquid or solid substance that have been emitted into the atmosphere and are present in a concentration high enough to be considered as harmful to the environment, or human, animal and plant health.

Air pollutants may be either emitted directly into the atmosphere so called “primary air pollutant” or formed within the atmosphere itself by reaction with other pollutants, “secondary air pollutants”.

Primary air pollutants are those which are emitted directly into the atmosphere from a source, such as factory chimney, exhaust pipe or through suspension of contaminated dust by the wind. Considering the mechanism of their appearance, it is possible to measure the amounts emitted at the source itself.

The share of primary pollutants in the atmosphere as well as their main type of sources is presented on the Figure 3 (EPA, May 2014).

Figure 3: Primary Pollutants Share



The main primary pollutants that are known to cause harm in high enough concentrations are the following:

- Carbon compounds: CO, CO<sub>2</sub>, CH<sub>4</sub> and VOCs,
- Nitrogen compounds: NO, N<sub>2</sub>O and NH<sub>3</sub>,
- Sulfur compounds: H<sub>2</sub>S and SO<sub>2</sub>,
- Halogen compounds: chlorides, fluorides and bromides,
- Particulate Matter (PM or aerosols), either in solid or liquid form,

Sources of primary pollutants are many. Natural sources of primary pollutants are volcanoes, fire, bacteria, viruses, pollens, blowing dust, etc. and they are accentuated by humans in recent history. However, the biggest contributions are causing sources created by humans.

After the emission, air pollutants in the atmosphere undergo dispersion and transportation, mainly due to meteorological conditions, chemical reactions and photochemical reactions. Thus, secondary air pollutants are formed. Because of this mode of formation, secondary pollutants cannot readily be included in emission inventories, although it is possible to estimate their formation rates.

The main secondary pollutants that are known to cause harm in high enough concentrations are the following:

- NO<sub>2</sub> and HNO<sub>3</sub> formed from NO,
- Ozone (O<sub>3</sub>) formed from photochemical reactions of nitrogen oxides and VOCs,
- Sulfuric acid droplets formed from SO<sub>2</sub> and nitric acid droplets formed from NO<sub>2</sub>
- Sulfates and nitrates aerosols, formed from reactions of sulfuric acid droplets and nitric acids droplets with NH<sub>3</sub>, respectively,
- Organic aerosols formed from VOCs in gas-to-particle reactions.

Another important distinction must be made in relation to the physical state of the pollutant. There are two categories: gas and particle. Gaseous air pollutants are those present as gases or vapors. They are readily taken into the human respiratory system and very often are precursors of adverse effects to human health. Gaseous air pollutants include NO<sub>2</sub>, SO<sub>2</sub>, CO, O<sub>3</sub>, NH<sub>4</sub>, VOCs, vapor phase of semi-volatile organic compound, etc. Particulate air pollutants mainly refer to fine particles in solid or liquid phase suspended in the atmosphere. Such particles can be either primary or secondary and cover a wide range of sizes.

Apart from the physical state it is also important to consider the geographical location and distribution of sources as well as their geographical scale (point, line or area sources). Depending primarily on the atmospheric lifetime of the specific pollutants, the local, regional and global scale of air pollution can be distinguished (WHO, May 2014).

## 1.4 Particulate Matter

Particulate matter, also known as particulate pollution or PM represents a complex heterogeneous mixture of liquid droplets and extremely small particles having diverse chemical and physical characteristics. It encompasses many different chemical components such as organic chemicals, metals, acids (such as nitrates and sulfates) and soil or dust particles, many of which have been specified as potential contributors to toxicity. Each of these components has multiple sources and each source generates multiple components. Sulfur dioxides and nitrogen oxides, known as secondary particles make up most of the fine particle pollution (Kelly and Fussell, 2012).

The size of particles is directly linked to their potential for causing health problems (Figure 4). Particles can be solid particles or liquid droplets which diameters are ranged from 0.1-50 micrometers. Those particles are called total suspended particles (TSP). Small particles, less than 10 micrometers in diameters cause the greatest problems, because they can get deep into lungs, and some may even reach bloodstream. Small particles which diameters are less than 10 micrometers are further divided into two major groups according to the size. Generally, inhalable coarse particles with diameter smaller than 10 micrometers and larger than 2.5 micrometers (PM<sub>10</sub>) and fine particles with diameters that are 2.5 micrometers and smaller, (PM<sub>2.5</sub>).

Figure 4: Particulate Matters Size (EPA, May 2014)



Inhalable coarse particles or PM<sub>10</sub>, are mainly mechanically produced by the break-up of larger solid particles. In urban areas, the coarse particles typically contain resuspended dust from roads and industrial activities, and biological material such as pollen grains and bacterial fragments. Typically, these particles also include wind-blown dust from agricultural activities, uncovered soil or mining operations. Near coasts, evaporation of sea spray can also produce large particles. Coarse particles may also be formed from the release of non-combustible materials in combustion processes, i.e. fly ash.

Fine particles, those smaller than 2.5 micrometers, are largely formed from gases. However, combustion processes may also generate primary particles in this size range. Generally, these particles originate as ultrafine particles (nuclei) produced by chemical reactions in the atmosphere, from various processing of metals, driving automobiles or burning plants.

Smaller particles are lighter and they stay in the air longer and travel further. PM<sub>10</sub> stay in the air for minutes or hours, while PM<sub>2.5</sub> can stay in the air for days or weeks. Regarding the travel distance, PM<sub>10</sub> particles can travel from few hundreds of meters up to 50km, still PM<sub>2.5</sub> can go even further, many hundreds of kilometers (Stephanou, 2012).

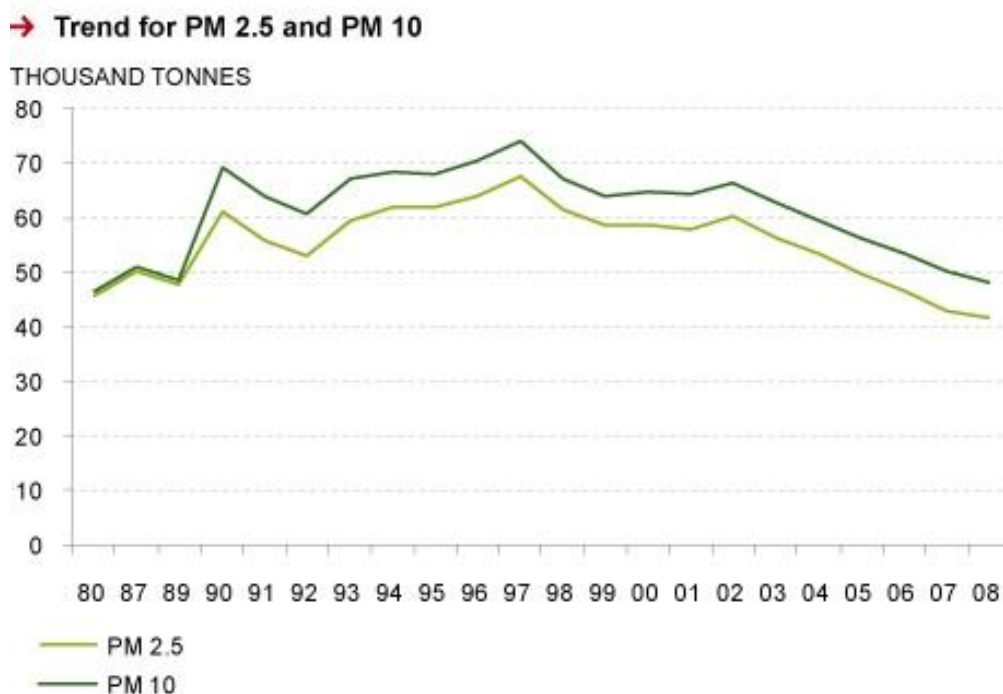
The particle matters may become dangerous to our health when we are exposed to it for a long time, and also when we breathe in a large amount of it. Apart from that, health effects can be acute or chronic. Acute health effects are characterized by sudden and severe exposure and rapid absorption of the substance. Normally, a single large exposure is involved and health effects are often reversible. On the contrary, chronic health effects are characterized by prolonged or repeated exposures over many days, months or years. Symptoms may not be immediately apparent. These kinds of effects are often irreversible (OSHA, May 2014).

Epidemiological studies have found a broad number of evidences on the association between exposure to air pollution and cardiovascular events. Scientific statement on particulate matter made by the first American Heart Association (AHA) concluded that short-term exposure to PM contributes to acute cardiovascular morbidity and mortality. Likewise long-term exposure may reduce life expectancy.

Notwithstanding, ambient air quality has been improved on the global scale in the past decade (Figure 5.), by following current policies for its abatement. But despite regulatory effort, fine particulate continues to be a matter of concern despite its falling trend. Protection of human health is further deteriorated considering the inability of scientists to establish a safe level of PM<sub>2.5</sub> below which it poses very little or no effect on human health. Furthermore, the effect of particulate matter on health is very complex since it may vary from one individual to another. It must be taken into consideration that not all individuals are equally vulnerable to air pollution health effects. Vulnerability could be

strictly linked to individual characteristics such as genetics, age, gender and the life style. For instance, low socioeconomic classes tend to be more vulnerable to adverse effects of air pollution considering their lower quality of life caused by other factors. Hence, analysts must calculate changes in health outcomes by taking into account that effect of pollution may be easily correlated with other elements that may be just as influential (Nazelle et al. 2011).

Figure 5: Global Trend of PM<sub>10</sub> and PM<sub>2.5</sub> (EEA, May 2014)



Exposure to PM can affect both your lungs and heart. PM<sub>2.5</sub> travels deeper into the lungs and since it is made of components which are more toxic (heavy metals), PM<sub>2.5</sub> can have worse health effects than the bigger PM<sub>10</sub>. Numerous scientific studies have associated particle pollution exposure to a variety of problems, including:

- Irregular heartbeat,
- Aggravated Asthma
- Lung damage (including decreased lung function and lifelong respiratory disease)
- Nonfatal heart attacks,
- Premature death with people with heart or lung disease
- Increased respiratory symptoms, such as irritation of the airways, coughing or shortness of breath (Koton et al. 2013).

On the other hand, epidemiological studies have shown that pollution acts synergistically with tobacco smoking, alcohol consumption and unhealthy diet to induce respiratory illness such as asthma, lung cancer and cardiovascular diseases.

However, there is a general agreement among scientists that fine particle matter (PM<sub>2.5</sub>) composition also plays a meaningful role in the health effects attributed to PM. Composition of PM may be more substantial than PM concentration alone in explaining health impacts. As evidence linking composition to health impacts emerges in the epidemiological and toxicological areas, it is becoming more urgent to distinguish which components or combination of components are the most harmful to human health. Toxicological studies suggest that several elements, including aluminum, silicon, vanadium, carbon-containing components and nickel are the most closely correlated to health impacts. However, many other elements have been implicated as well. There are no PM components for which there is unequivocal evidence of zero health impact (Rohr and Wyzga, 2012).

Carbonaceous classes, elemental (EC) and organic carbon (OC) are well known contributors to the atmospheric particulate matter at a global level, and also very frequently dominant contributors to the fine particulate matter mass. EC is derived from incomplete combustion of carbon-based materials and fuels and is present in primary form in the nature, while on the other hand OC may be directly released into the atmosphere or produced by way of secondary gas-to-particle conversion process. Along with the adverse health effects which are correlated with fine particulate matter mass exposure, carbonaceous species inflict very serious health effects. Considering the fact that EC is mainly considered as inert and that combustion process is deriving EC, coated by organic matters as PAHs, which are widely known to have carcinogenic and mutagenic properties and to cause serious health risks (Lonati et al. 2007).

Knowledge about carbonaceous portions in fine particulate matters and especially their portioning between primary and secondary origin may be used very efficiently in conveying new air quality plans and targets (Lonati et al. 2007).

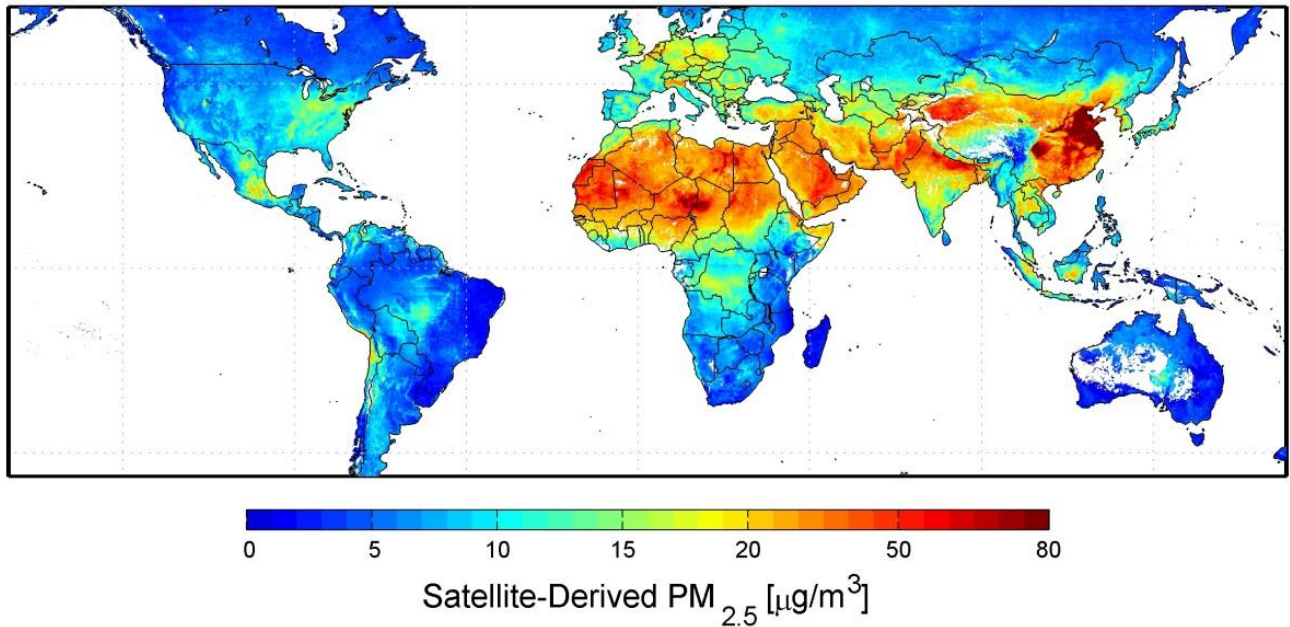
The particles are causing many serious problems both to humans and ecosystem. Hence it is vital to determine the number, morphology and size distribution of these particles so as to reflect their intensity and harmful effects on human health. Hence, global distribution of PM<sub>2.5</sub>, averaged over 2001-2006 is presented on the figure 6.

Identifying and quantifying the influences of specific components or source-related mixtures on measures of health-related impact represent one of the most challenging areas of environmental health research. Current knowledge does not allow precise quantification of the health effects of PM emission especially because particles interact with other co-pollutants. Therefore, a carefully targeted program research, including more refined approaches is needed to enhance our understanding of the relative toxicity



of particles. This approach will facilitate development of abatement policies, more effective pollution control measures and a reduction in the rate of health problems caused by particulate pollution.

**Figure 6: Global Satellite Deriver Map of PM<sub>2.5</sub>, averaged over 2001-2006 (WUWT, May 2014)**



Besides health problems to humans, particulate pollution also drives other damages. For instance, fine particles PM<sub>2.5</sub> are the main cause of reduced visibility (haze). In a form of acid rain they can damage stone and other materials, including culturally important statues and monuments. Acid rain refers to a mixture of wet and dry deposition from the atmosphere containing higher than normal amounts of nitric and sulfide acids. It can appear in a form of rain, snow, fog and tiny bits of dry material that settle on Earth. Origin of formation of such rains lies in the PM either from natural sources or from anthropogenic sources. Considering the fact that particles can be carried over a long distances by wind and then settled on ground or water, this effect can make also lakes and streams acidic. By settling into the other mediums, particles may also change their nutrient balance, damage sensitive forests and farm crops, deplete nutrients in soil and lastly affect the diversity of ecosystem (Liang, 2013).

Apportionment of pollution sources should be the main interest in development of current strategies and regulations of the environment. It is important to acquire as many information as possible about the sources of pollution we are exposed to, as well as characteristics and concentration of those sources. Hence, if a specific source that contributes to air pollution in a given area is known, than strategies can be planned and implemented in a way that reduces the impact of those sources. Additionally, by knowing the type of the sources that are contributing the most to the air pollution,

awareness of the people may be raised to a higher level, thus they can make some modifications in their life style and habits.

## 1.5 Trends and Projections

Europe needs more resource efficient, greener and competitive economy. Consumers who appreciate resource efficiency, new green technology and smart inventions toward sustainability can create new economic opportunities. By developing cleaner and more efficient energy, this green industry can create new jobs and this approach can as well reduce Europe's import of oil and gas and improve its position on the energy market. With development of such technologies, it will not be achieved just improvement towards the environmental sustainability but at the same time and economic recovery (ESPON, May 2014).

Accordingly, some of the main trends present nowadays are as follows:

- Waste treatment processes in EU have improved remarkably since 2000. Landfilling represents one of least environmental-friendly technics for managing waste. Accordingly incinerators started gradually replacing them with a greater frequency by applying composting and recycling techniques. About 40% of municipal waste was recycled or composted. However, there are huge fluctuations in techniques used in some countries in Europe. For instance, in Bulgaria, Croatia and Romania more than 90% of waste was landfilled, where on contrary in Germany, the Netherlands and Sweden this approach was below 1%.
- Current estimates are showing that the extent of the effects of ozone and fine particle pollutants on life expectancy is in the order of several tens to hundreds of thousands of premature deaths per year in Europe (WHO, 2006). Despite the fact emissions of air pollutants are generally declining, many countries are not jet on a track towards EU targets and air quality limit values for PM<sub>10</sub>, PM<sub>2.5</sub> and NO<sub>2</sub>. Even if the emission reduction targets are met, health impacts are still likely to occur. However, this appears partly due to background levels and natural sources of these compounds, which is impossible to address from a European perspective only.
- Energy consumption in Europe peaked in 2005 and has been declining since. This trend decelerated slightly until 2010, partly due to the economic crisis and limited economic recovery in 2010. This result is clearly due to implementation of energy efficiency and renewable energy policies, where the economic crisis and structural changes also played a significant role in the most recent trends taking also into account milder winters. More recently, primary energy consumption in Europe was 14.4 higher than the 2020's set target. The most significant decrease

occurred in Germany, France, Spain, Italy and the United Kingdom. The economic crisis was more pronounced in these countries especially in the industry and transport sectors. Also, the switch of fuels played a major role in this decrement. Lastly, in 2011, the final energy consumption in EU-28 was only 2.4% higher than the 2020's set target.

- On the other hand, there is a rapid expansion on renewable energies, particularly in electricity sector (Table 2). Energy generated from biomass, wind, solar energy and the Earth's heat are replacing share of fossil fuels in the final energy demand in the EU. Between 2005 and 2011 all member states of the EU have increased their renewable energy share. While it is asserted the greatest expansion in wind and solar energy, however contribution of biomass represents by far the largest amount. In conclusion, renewable energy sources are covering a fifth of gross power generation in 2011.
- Despite the fact the gap in CO<sub>2</sub> emissions per capita narrowed between the EU and developing countries from 2000 to 2011, the CO<sub>2</sub> emissions remained at 7.4 tones per capita in the EU, which is 2.6 times greater than the developing country average of 2.9 tones per capita. This narrowing of gap occurred primarily as a result of increasing emissions from developing countries and on the other hand financial crisis which led to reduction of CO<sub>2</sub> emissions per capita in the EU (Eurostat, May 2014).

**Table 2: Contribution by Renewable Energy Carriers (electricity, heating/cooling, transport) (EEA, May 2014)**

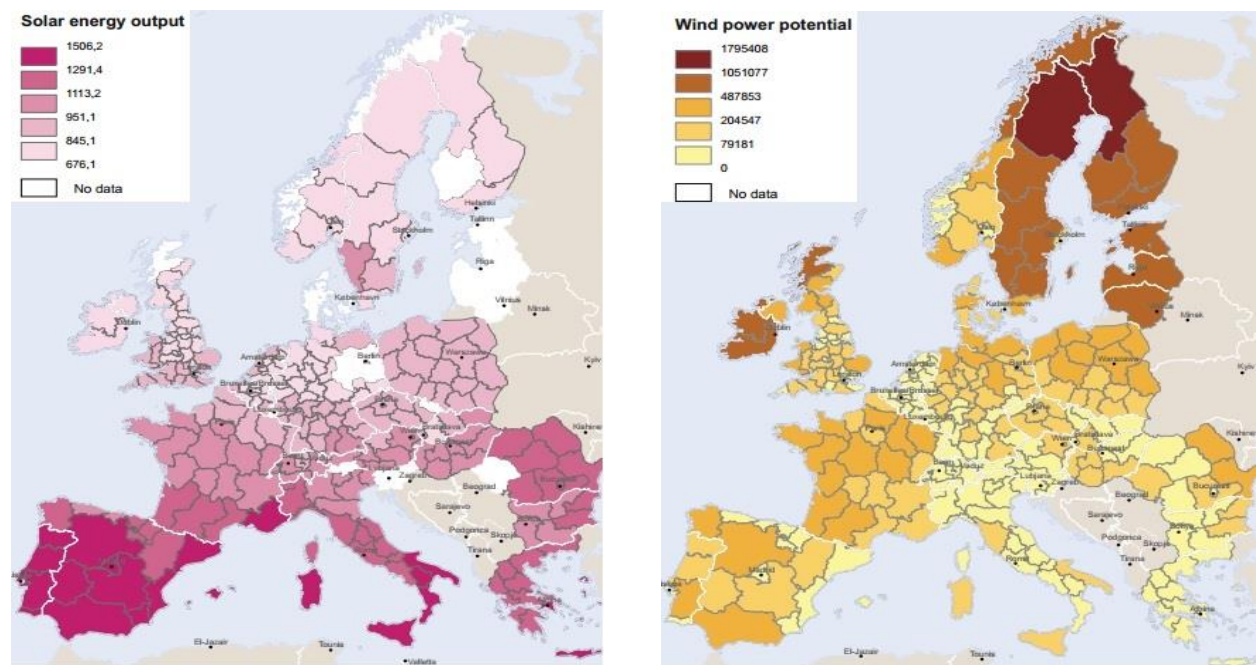
Year	Energy (Mtoe)				Share (%)
	2005	2010	2011	2020	2020
RES-E	41.4	55.9	60.7	104.2	42
RES-H/C	58.9	78.4	76.7	111.5	46
RES-T(including biofuels)	1.0 (4.2)	10.5 (14.4)	11.5 (15.0)	29.5	12
<b>Total RES (including biofuels)</b>	100.3 (103.4)	143.6 (147.6)	147.2 (151.2)	245.1	100

Current estimates are indicating that EU is having 7.7% of the world's population and contains 9.5% of the world's biocapacity. Ability of the system to generate biological materials and to absorb waste materials produced by humans accounting current management approaches and technologies is called biocapacity. Despite having above average biocapacity with respect to its population, EU generates 16% of the world's ecological footprint. What is more, EU' development strongly depends on the ecological reserves in other parts of the world.

Nevertheless, some scientists believe that large-scale biomass plants could accelerate deforestation or endanger local biodiversity. These observations are not an argument against evolution of biomass as a new renewable energy industry. Rather they want to emphasize the need for an integrated approach to territorial development.

Considering other renewable sources of energy, mainly solar and wind, it can be said that great expansion is currently being made and still there is a lot of potential in these fields. The map of Europe with yearly total yield of estimated solar electricity generation (kWh) and production potential of wind power stations, taking into account environmental and other constraints are presented on figure 7 and figure 8. From the maps it can be seen that the North Europe dominates in a potential for wind production, while Southern Mediterranean Europe dominates in the high potential for a generation of energy using Sun energy.

**Figure 7 and 8: Yearly Total Yield of Estimated Solar Electricity Generation (kWh) and Production Potential of Wind Power Stations. (ESPON, May 2014)**



The EU-15 has a common objective to be achieved collectively under the Kyoto Protocol. This protocol in an international agreement linked to the United Nations Framework Convention on Climate Change, which sets binding obligations on countries to reduce emissions of greenhouse gasses. It sets emission limitations and reduction targets for each EU-15 Member State. Each of the targets corresponds to an emission budget (“Kyoto units”) for the first commitment period (2008-2012) of the Kyoto Protocol. Nearly all Member States and all other European Environment Agency (EEA) countries achieved targets set by Kyoto protocol by the end of the Kyoto Protocol’s first

commitment period, i.e. the target was set to 8% of reduction for the period of 2008-2012 compared to base-year levels under the Kyoto Protocol and EU-15 have reached reduction of 12.2%.

However, as in previous years, Italy remains considered off track towards the target, mainly due to lack of information on its planned use of flexible mechanisms. Italy did not put a threshold on the use of flexible mechanisms in its national climate change strategy, but administrative arrangements are being taken for purchases (EEA, 2013).

Europe 2020 calls for a perception of structural and technological changes in order to move to low carbon, resource efficient and climate resilient economy by 2050. This progression will enable Europe to meet its emission reduction targets. It will include new business growth to sustain Europe's leading role in green technologies on the international market, but also disease prevention and response as well as adjustment measures based on more efficient use of resources.

The EU has set 5 targets to be reached until 2020. These targets are mainly focused on improvement of employment rates, greenhouse gases reduction, poverty reduction and enhancement of educational attainment and health. Likewise, "20/20/20" represents the triple objective for 2020. These targets are endorsed by the European Council in 2007 and implemented through the EU's 2009 climate and energy package and the 2012 Energy Efficient Directive, and it focuses on:

- A 20% increase in energy efficiency
- A reduction in greenhouse gas emissions by at least 20% compared to 1990 levels.
- To develop renewable energy resources so that they can account for 20% of total energy consumption (ESPON, May 2014).

The time scale is fundamental to implementing methods for a higher sustainability. Europe 2020 represents a short-term time horizon with respect to economic development and short-term horizon from the perspective of Europe's sustainable development. Accordingly, it is very important to look further into the future and not just into the meeting the current targets. This is fundamental as environmental impacts are long lasting, climate change is a long-term process and patterns of urban areas, energy and transport networks need time to adapt.

The territorial impacts on the scenario are under change over time. The primary impact is caused by the metropolitan regions, mainly in Western Europe, thus the greatest expectations in investments in new technologies are expected to be made in this region. However, in the later phase of the scenario there is a diffusion of these ideas and a more polycentric pattern of growth. Nonetheless, it is important to identify that such

approach carries more challenges for some regions than for others. South and East European cities might be more confronted by sustainable issues, as they have less advanced public transport systems, more polluting cars, buses and trucks, lower amount of green areas, old technology etc.

## **1.6 Need for Source Apportionment**

Abatement of pollution at its source represents one of the main principles of the Thematic Strategy on Air Pollution. Information on pollutant sources is essential to the design of air quality policies. Therefore, source apportionment is required for the implementation of the Air Quality Directives (Dir.2008/50/EC and Dir.2004/107/EC). For instance, the true pollution source information is required in order to identify whether possible overcoming are due to natural sources or to artificial ones, set the limit values of pollutants, preparing air quality plans, quantifying trans-boundary pollution, and informing the public (Belis et al., 2014).

Particulate matter (PM) is one of the main pollutants exceeding the ambient standards for air quality in Europe. Majority of suspended particles are too small to be seen with a human vision. Even if they are identified with some magnification technique, their origin very often remains unknown. A variety of sources contribute to a specific event and the proportions of these contributions change from event to event, driven by many influencers. Some particles preserve the forms in which they were originally emitted, but others are created from emitted gases through chemical reactions, or they undergo chemical transformations that change their chemical and physical characteristics. These facts has let to numerous studies focusing upon its complex composition, toxicology and source attribution (Watson and Chow, 2007).

One of the most powerful tools for the formulation of abatement policies and verification of their effectiveness is particulate matter source apportionment by the combination of chemical and statistical analysis. Source apportionment is the practice of deriving information about pollution sources and the amount they contribute to ambient air pollution levels (Frao et al. 2013).

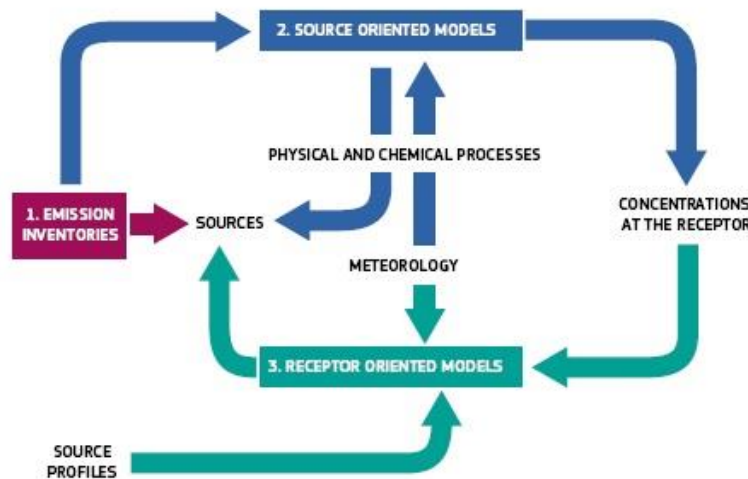
Different approaches are used to determine and quantify the impacts of air pollution sources on air quality. Commonly used techniques are:

- Explorative methods
- Emission inventories
- Inverse modelling

- Artificial neural networks
- Lagrangian models
- Gaussian models
- Eulerian models
- Receptor models

However, contributions to PM levels can be highlighted through two source apportionment approaches, receptor and source models. These models are complementary rather than competitive. Receptor models are centered towards the properties of the ambient environment at the point of impact, while the source-oriented dispersion models take into consideration transport, dilution and other processes that take place between the source and the sampling site. Each has strengths and weaknesses that compensate for the other (Figure 8). Both types of models can and should be used in an air quality source assessment of outdoor and indoor air (Bove et al. 2014).

Figure 9: Schematic Representation of the Different Methods for Source Identification (JCE report)



Receptor-based models use chemical measurements (chemical composition, particle size and concentration variation in space and time) at an individual monitoring site (the receptor) to calculate relative contributions from major sources to the pollution at that site (EPA, May 2014). Receptor modelling cannot predict future air quality but, instead, looks at past data collected at one site over a specific time period to determine the sources to that site. Receptors can be stationed indoor or outdoor or they can be mobile samplers. This model reconstructs the contribution of each source by processing series of PM compositional values that are measured at specific receptor's sites i.e. multivariate analysis is used to solve a mass balance equation to identify and allocate sources of the PM in the atmosphere. These tools have the advantage of providing

information derived from real-world measurements, including estimations of output uncertainty. However, these models can be unsuccessful with reactive species and may perform better in areas relatively closer to the sources (Bove et al. 2014).

Receptor models have been widely used over past three decades to apportion ambient concentrations to sources. Among these, the Chemical Mass Balance (CMB), Principal Component Analysis (PCA) and Positive Matrix Factorization (PMF) methods are the most frequently used. CMB can be used if sources and emission profiles of PM are known “a priori”. However, a detailed knowledge of sources and emissions is not always available; in these cases it is preferable to use multivariate models like PCA and PMF, which attempt to apportion the sources on the basis of the internal correlations at the receptor site.

The main output from these models is an estimate of the contributions from each source to the air pollution at that site. However, the reliability of receptor model outputs depends on appropriate data collection, in terms of data capture and kind of chemical species, and proper expression of uncertainty in the input data. This represents an aspect particularly relevant in PMF, which scales data on the basis of their uncertainty. In addition, determining the number of relevant sources and establishing the correspondence between factors and sources still appear as critical steps. Results from these models are important for scientifically justifying priorities and observing trends. Moreover, this scientific information helps air quality modelers as well as policy and decision makers (EC, June 2014).

An alternative to the statistical data analyses is given by simpler models based purely on chemical analysis of dominant PM components and is called source models. Source models estimate concentrations at receptor coming from different source emissions and being influenced by meteorological measurements. Basically, chemical determinations are individually summed up in order to obtain a mass closure and then grouped to determine the macro-sources of PM. To enhance the selectivity of the elements as source tracers, a size fractionation of PM can be performed, as it is well known that fine particles ( $< 2.5 \mu\text{m}$ ) are mainly emitted from combustion sources and coarse particles ( $> 2.5 \mu\text{m}$ ) are generated from mechanical-abrasive processes (Watson et al., 2002).

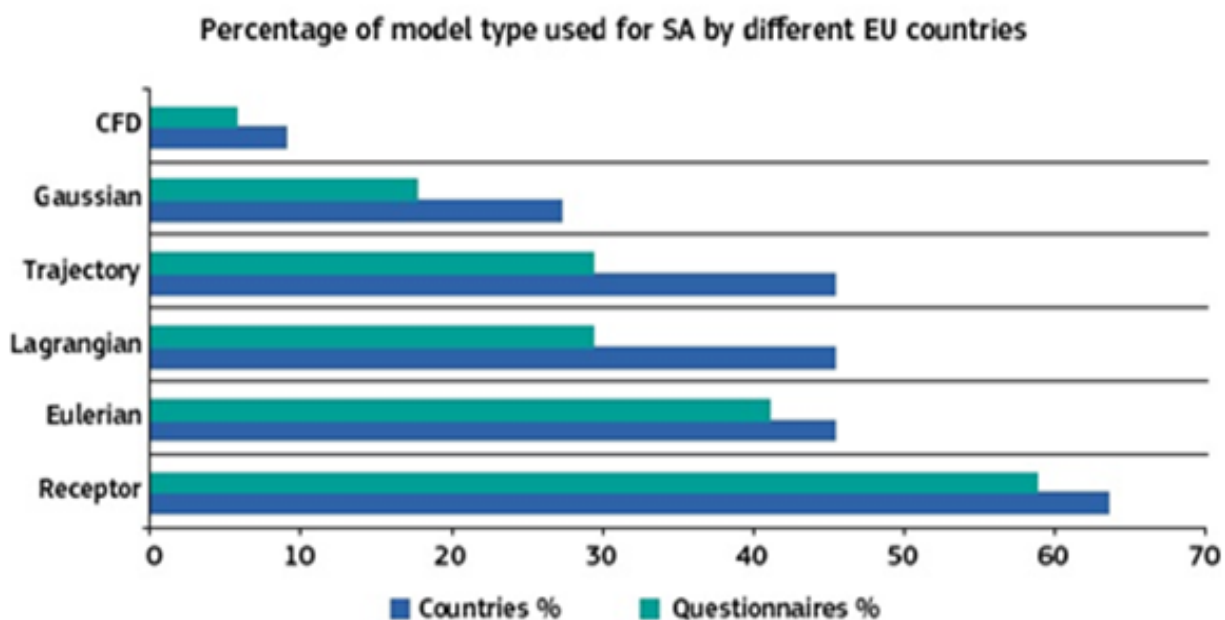
Within the activities of the Forum for Air Quality Modelling in Europe (FAIRMODE) group on “Contribution of natural sources and source apportionment”, few surveys were implemented with the focus on the type and frequency of model used for source apportionment in Europe. When examining the received information on the most recent survey (Fragkou et al. 2012) it becomes obvious that the different tools for source identification in Europe ranged from less than 20% for Gaussian models to almost 60% for receptor models used in Europe (Figure 9). Furthermore, among all technologies



used in Europe, the most used models in Italy are Receptor (CMB, PMF<sub>2</sub>, PMF<sub>3</sub>) and Dispersion-Eulerian (FARM, CAMx/PSAT) models.

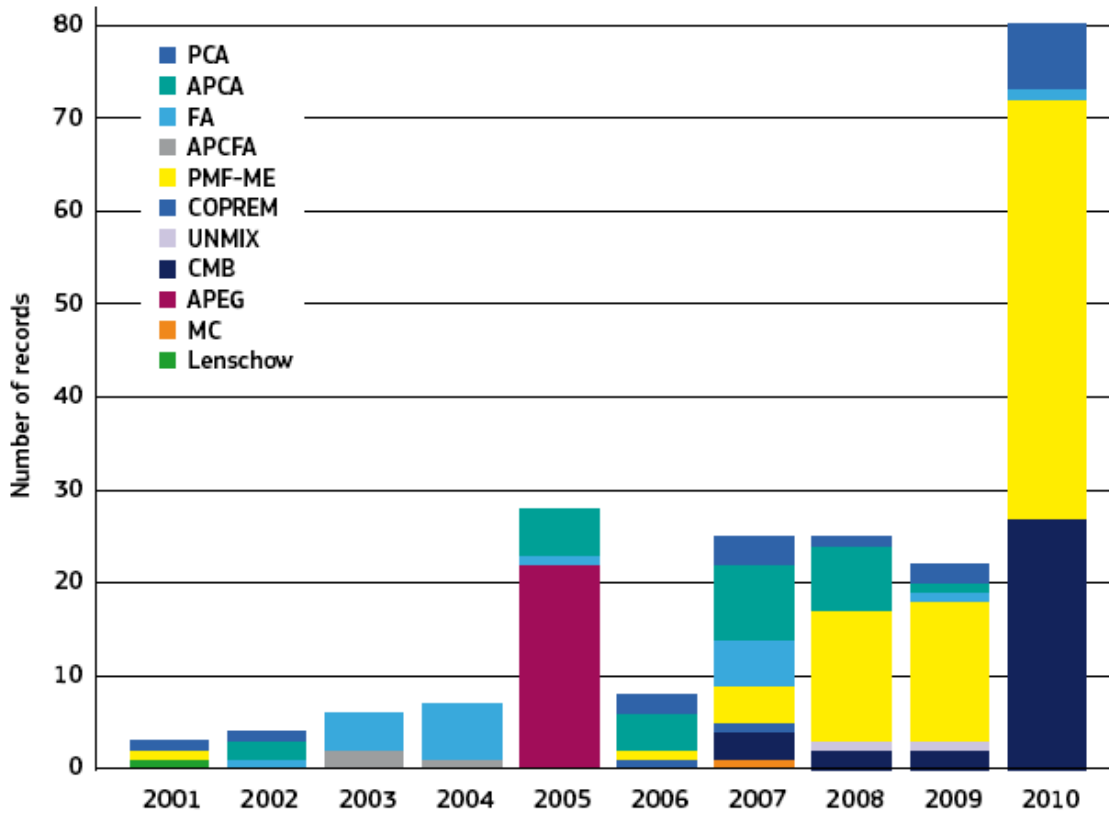
The Receptor models for apportionment of sources and their pollutants are widely recognized from all EU countries. Such a large number of Receptor model's presence may be induced for its low computational intensity, source estimation at the urban and regional scales which are independent from emission inventories and meteorological data preprocessing (Fragkou et al., 2012).

**Figure 10: Percentage of Model Types used for Source Apportionment by Different EU Countries (Fragkou et al. 2012)**



Previous study by Viana and co-authors implemented an analysis of source apportionment studies in Europe from 1987 to 2007. According to this study, PCS was the most frequently used model up to 2005 by having 30% of study and PMF and CMB were respectively 8% and 7%. In following years, an increase in the usage of PMF (13%) and the mass balance analysis of chemical components (19%) occurred (Viana et al. 2008). In contrast with the tendency observed between 1987 and 2005, the majority of the studies were performed by using PMF and CMD models in the period of 2001 and 2010 (Figure 10).

Figure 11: Time Trend of RM Studies in Europe between 2001 and 2010 (Belis et al. 2014)



The obvious advantage of the PMF models over CMB relates the fact that related software is widely available and detailed information on the sources and source profiles is not required. On the contrary, CMB model requires expanded data on pollution sources prior to source apportionment.

In order to compare all the source apportionment results and to derive useful conclusion, sources have been derived into six major categories, representing the most frequent ones:

- Sea/Road Salt
- Crustal/Mineral Dust
- Secondary Inorganic Aerosol (SIA)
- Traffic
- Point Sources
- Biomass Burning (Belis et al., 2014).

Considering the target metric, until 2005, PM<sub>10</sub> was, on average, the preferred target, by having more than 46% of the publication reported, and it was followed by 33% of PM<sub>2.5</sub>.

While other smaller size fractions  $PM_2$ ,  $PM_1$ ,  $PM_{0.1}$  had smaller share. However, different trend occurred starting from 2006, when  $PM_{2.5}$  took over the lead in the share by having 38% of the new studies and on contrary only 29% focused on  $PM_{10}$ . Thus, this new trend shows a change of focus in source apportionment studies in Europe. This new order may be due to the stronger recent evidences on the adverse effects of fine particles on health with comparison to coarser particles (Viana et al. 2008).

The majority of the studies were implemented in urban areas (53% of the studies) while industrial or kerbside sites represented respectively 11% and 20% of the studies (Viana et al. 2008).

The results of all surveys demonstrate the simultaneous use of different modelling tools and methods for identification and attribution of sources at investigated receptor sites, as the applicable methodology in order to combine the advantages and reduce the constrains of the individual model components. Moreover, the trends in usage are changing rapidly due to new discoveries on adverse effects of particulate matter mainly on the human health.

## 1.7 Scope of the Work

The fundamental objective of this work is to give contribution to better understanding of the nature of PM and to quantify relative significance of diverse emission sources and enable decision makers to convey efficient air quality remediation plans. Thus, the scope of this work is to perform a PM source apportionment by means of the PMF model on the PM data set from Milan<sup>2</sup>.

In the recent past years during PM monitoring campaigns the data set is acquired with an aim to define the composition and contributions of various species to  $PM_{2.5}$  bulk mass in Milan. Thus, this work will contribute to enrichment of the current records for the specified area.

The main focus is on  $PM_{2.5}$ , since recent air quality standards for PM inquired by European Union, studies also  $PM_{2.5}$ , besides  $PM_{10}$ , where were specified both, concentration limits and an exposure reduction target. The second reason for centralizing on  $PM_{2.5}$  is that anthropogenic carbon forms are almost completely included in fine PM (Lonati et al., 2007).

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<sup>2</sup> The city of Milan is situated in the Lombardy region and is surrounded by many industrial zones; Furthermore, the metropolitan area of the city has a population of over 4 million people.

## 2 Model Application and Material

### 2.1 PMF Application

PMF represent a recent development in the data analysis technique's sphere, which is called factor analysis. The main problem in these techniques is to resolve the identity and contribution of components in an unknown mixture. PMF is particularly applicable in the projects that use environmental data, for the following reasons:

1. It integrates the uncertainties of variables which are often associated with measurements of environmental samples
2. It drives all the values in a solution profiles and contributions to be non-negative, which represents more realistic approach than previously used models

PMF is being used to categorize and apportion sources of airborne PM, by collecting data in numerous locations in the World. Mainly the collection locations are being focused in the urban areas, since there is the biggest need for the knowledge of the ambient air components. Likewise, classification of PM is set as  $PM < 1\mu m$  (PM1),  $< 1PM < 2.5\ \mu m$  (PM2.5) and  $PM < 10\ \mu m$  (PM10). This information is used for categorization of the PMF case studies with respect to PM.

These data sets are principally used to identify profiles and contributions of PM from primary sources; for instance, motor vehicles, residential and industrial fuel consumption, biomass burning, soil dust and sea salt. Likewise, secondary sources, such as atmospheric oxidation of sulfate and nitrate and heterogeneous gas-to-particle conversion reactions on soil dust surfaces, are as well subject to the PMF application. (Prakash V. Bhave, et. al. 2007)

After consideration of the primary use of PMF, there are many more fields of application for this model, and many new ones are being under feasibility study.

Chemical composition of soil samples are multivariate in nature and hence represent the ideal data for the multivariate factor analytical techniques (PMF) and for its approximation. Recent enhancements of the model led to the increased application within this field of study. The reason lies in the fact that previous approaches used for the analysis of soil datasets, did not rely strongly on physically significant assumptions. By combining results from PMD model with geostatistical approach, it was possible to successfully determine the main sources of the combined organic and heavy metal contamination. (S. Vaccaro et.al. 2007)

Another combination of PMF with other approaches is used. Quantification of the diesel- and gasoline-powered motor vehicle emissions can be identified by merging results

from PMF receptor modelling and on-road measurements captured by a mobile laboratory. By obtaining firstly source profiles from the PMF, the calculation of fuel-based emission factors for each type of the exhaust is possible. (D.A. Thornhill, et.al. 2010)

On the other hand, PMF application within the research and development facility emissions is very convenient. The processes are varying and because of that the nature of research and number of chemicals are being changed rapidly, and PMF was the ideal match since it was able to identify the biggest number of source-related factors, while other approaches did not achieve such a good results. PMF is able to accept the boundaries with little reduction in model fit. (M. Y. Ballinger and T.V. Larson, 2014)

Eco-efficiency indicators are very significant tool if researchers want to create physically meaningful information to policy makers. PMF is limiting its results to be non-negative, and with it two important advantages over traditional factor analysis are achieved. Firstly, the rotational ambiguity of the solution space is reduced. Secondly, all the results are guaranteed to be physically meaningful. (J. Wu et.al. 2012)

PMF, however, finds its place in a very technical area of studies as well. Time-resolved optical waveguide absorption spectroscopy (OWAS) is a technique used for the investigation of kinetics at the solid/liquid interface of dyes, pigments, fluorescent molecules, quantum dots, metallic nanoparticles, and proteins with chromophores. The application of PMF to these techniques is quite recent, but it is already proved that it prevents the negative factors from occurring, avoids contradicting physical reality and makes factors more easily interpretable. (P. Liu et.al. 2013)

On the contrary of PMF extensive use, there are significant fluctuations in the procedure process for the source apportionment. This procedure may be separated into three broad steps:

1. Preparation of data to be modeled
2. Processing the data with PMF with an aim to develop a realistic and robust solution
3. Interpretation of the solution

Some specific decision making are needed in the above mentioned steps, towards the choice of data uncertainties' set, selection of factors and treatment of outliers.

## 2.2 PMF within Europe

Danube River represents the second longest river in the Europe. Its spring starts in the Germany's Black forest and it flows until its delta on the Black Sea. Considering its size and impact that generates, the chemical composition of the river and its tributaries should be well identifiable. In the past, some monitoring programs were performed in various parts of its drainage basin, taking into consideration its tributaries as well, with an aim to quantify the micro-pollutants level in the river. However, as new methods were developed, PMF was used with an objective to identify both natural and anthropogenic sources affecting Danube basin, as well as origin of heavy metals and other possible sources affecting the sediment creation. With applied PMF method, the spatial distribution of resulting sources was used to identify the role of the tributaries as potential sources of pollution. Results showed that the majority of tributaries are influenced by the anthropogenic sources. For instance, Velika Morava River has very high concentration of metal in sediments which can be influenced by the mining activity in the catchment area. On the other hand, the Sava tributary showed elevated values of mercury, probably in association with old refinery activities and chemical industry. In conclusion, the PMF application identified one anthropogenic parameter, which could be linked to different anthropogenic activities depending on the location along the Danube River: municipal and industrial discharge and mining activity (S.Comero et. Al. 2014).

One of the hotspot areas in Europe with high concentration of particulate matter which is having a lot of problems in meeting all standards for current legislation of PM<sub>2.5</sub> is the Netherlands. For the sake of better understanding of current levels, composition and distribution and origin of PM<sub>2.5</sub> in the ambient air, one-year measurement campaign was run in the five locations in the Netherlands area. PMF was used as the main tool to understand and categorize the most relevant source contributions and their spatial variability in PM<sub>2.5</sub>. Wind direction was as well incorporated into the study of the results from PMF with an aim to identify more accurately the possible locations of the identified sources. (D. Mooibroek et.al. 2011)

To have a wide outlook of PM<sub>1</sub><sup>3</sup> mass concentration and chemical composition of sub-micron sized aerosols, two measurement campaigns were performed in three towns in Italy, Milan, Genoa and Florence. Every town was having different characteristics with respect to their orography, extension, population and emission sources. This campaign represents first large-scale investigation on PM<sub>1</sub> in Italy and likely in Europe. The aim of this research campaign was to identify major sources of PM<sub>1</sub>, and to estimate their contribution to mass concentration. After running the PMF model on the collected data sets, it is identified that during the wintertime, the highest concentrations of PM<sub>1</sub> were in

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<sup>3</sup> PM<sub>1</sub>, particulate matter with aerodynamic diameter smaller than 1  $\mu$  m,

Milan, due to the high loading of pollutants and the atmospheric stability. Since, Po valley where Milan is situated, has peculiar meteorological conditions and very heavy emissions of pollutants from many different sources, it represents the most critical area in Europe with respect to limit values exceedances. However, during summer time PM<sub>1</sub> concentration significantly decreased in Milan. As the reason of this reduction, slower average wind speeds and mixing layer heights were crucial influencers. On the contrary, other two cities were showing completely different results since conditions in those areas differ significantly with respect one in Milan area. Lastly source identification was carried out with the help of available literature source profiles for PM<sub>1</sub> fine fractions, by looking at source contribution time series and by taking into consideration explained variation values. (R. Vecchi et. al. 2008).

Data set collected in the urban area around Elche in southern Spain starting from December 2004 until November 2005 was analyzed with PMF in order to estimate sources profiles and their mass contributions. After running the model, six sources were identified. However, very important it to mention that with the PMF it was possible to distinguish Saharan dust sources from local dust sources, and to quantitatively estimate the contributions of these two sources. (J.Nicolás et.al.2008).

Another sampling was used as well in Spain, Zaragoza city, but this time the collection was oriented towards the PM<sub>2.5</sub> PAH<sup>4</sup> substances. The aim of this project was to identify and quantify potential PAH pollution sources. As the origin of the PAH is in the fossil fuels, it does not surprise the fact that the most influencing sources were found in coal combustion, vehicular emissions, stationary emissions and unburned/evaporative emissions. In an evaluation of the results of the PMF model run, some different patterns were identified and same were studied for the identification of the potential negative impact on the human health. Above all, lifetime cancer risk exceedances were examined for both, warm and cold seasons. (M. S. Callén, A. Iturmendi, J.M. López, 2014)

The similar study was performed in UK National Network between 2002 and 2006. The goal of the study was apportionment of PAH sources as well, since they are currently generating a great deal of interest, mainly because of their high toxicity and carcinogenicity. The project was incorporating 14 urban sites which were known for the vast impact towards the creation of PAHs. (E. Jang, et. al. 2013).

After considering all above mentioned case studies, it can be concluded that PMF is widely used within Europe and for many different purposes. For instance, in the first project, PMF identified in a very complex system, major influencing sources to the

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<sup>4</sup> PAH – Polycyclic aromatic hydrocarbons are organic compounds containing only carbon and hydrogen that are composed of multiple aromatic rings. PAHs are found in fossil fuels when an incomplete combustion occurs because of insufficient oxygen. Critically, PAHs have been identified as carcinogenic and mutagenic and are considered as pollutants of concern for the potential adverse health impacts.

Danube River sediments. In another project conveyed in the Netherlands, the PM<sub>2.5</sub> was examined with PMF, and relevant source contributions and their spatial variability were categorized. PM<sub>2.5</sub> was under study of PMF but this time direction was toward identification of the PAH pollution sources. The studies were performed both in Spain and UK. Taking into account particulate matters, another project was conveyed, however this time PM<sub>1</sub> was studied, and the project took place in Italy. Another use of PMF was to differentiate types of sand in the Spanish south regions, thus PM<sub>10</sub>. Accordingly, it may be concluded that PMF represents very popular method for the identification of many different types of PMs in the vast range of European territories, and the trend of its use is in increasing phase.

### 2.3 PMF Model

Receptor models represent mathematical methods for quantifying and qualifying the contribution of sources to observed samples, based on the structure or fingerprints of the sources. The structure of the source is identified by using analytical approaches for the media and fundamental species or consolidation of species is needed to separate influences. A composition data set can be shown as a data matrix  $X$  of  $i$  by  $j$  dimensions, in which  $i$  is the number of samples and  $j$  represents number of chemical species that were measured, with uncertainties  $u$ . The aim of the receptor model is to resolve chemical mass balance (CMB) between measured source profiles and species concentration. The sources profiles are described with number of factors  $p$ , the species profile  $f$  of each source, and the amount of mass  $g$  which is created by influence of each factor to each individual sample. The CMB equation can be resolved with many models including 3 models that EPA has developed. In this work, EPA Positive Matrix Factorization will be used.

PMF represents a multivariate factor analysis program that decomposes a matrix of identified data set into two matrices, factor contributions (G) and factor profiles (F). These factor profiles needs to be studied by the user in order to allocate the source types that may be influencing to the same cluster, by using quantified source profile information and emissions of discharge registers.

Results are created using the boundary condition that no sample can have significantly negative source influence. PMD combines both user-identified uncertainties correlated to the sample data to weight individual points and sample concentration. This characteristic permits analyst to include confidence in the measurement process.

PMF model obliges several repetitions of the fundamental Multi-linear Engine (ME) with an aim to categorize the most optimal factor contributions and profiles. ME is created to



resolve the PMF problem by aggregating two steps. Firstly, the analyst creates a table which specifies the PMF model. Thereafter, a programmed secondary model reads the previously created parameter table and computes the solution. The best practice is to iterate the model approximately 20 times for the development of a solution and 100 times for the creation of the final solution, however different starting point should be used every time.

Variability due to chemical transformation or method fluctuations can influence considerably by causing significant differences in factor profiles among PMF runs. The analyst needs to identify all the error estimates in PMF to comprehend the strength of the model outcomes.

Lastly, PMF needs a data set containing a number of parameters which are measured across various samples. Usually, PMF is used on diverse PM<sub>2.5</sub> data sets containing from 10 to 20 species over 100 samples. An uncertainty value that is assigned to each specie and sample should be estimated by using the additional uncertainty data set. This data set is designed using propagated uncertainties or other available information such as collocated sampling precision. (EPA Positive Matrix Factorization (PMF) 5.0. Fundamentals and User Guide)

### **2.3.1. Comparison to PMF 3.0**

The PMF Model Development Quality Assurance Project Plan has been created to collect comments and suggestions from users. After completion of the project, very useful comment were used to improve the problematic features in PMF 3.0 and to create new tools for the improvement of the results' accuracy and error estimation; thus to develop version 5.0.

The main difference between PMF 5.0 and PMF 3.0 is in added two key tools. Firstly, two additional error estimation methods and secondly, source contribution and profile boundaries are added in the newest 5.0 version. Likewise, many other features have been added as well to enhance the software's friendly use, e.g. ability to read in multiple site data, etc (EPA Positive Matrix Factorization 5.0. Fundamentals and User Guide).

## 2.4 Input data

PMF requires two types of input data: 1. Sample species concentration values and 2. sample species uncertainty values or parameters for calculating uncertainty. Type of the files that EPA PMF admits are comma-separated values (.csv), tab-delimited (.txt) and Excel Workbook (.xls or .xlsx) files. The input process may be performed simply by writing the file path into the “data file” box or by browsing to the particular file. However, in some cases the file may contain multiple worksheets and user will be asked to choose only one to be run by model. The concentration file is determined by species as columns and sample numbers or dates as rows, including headers for each of them (Figure11). The accurate forms for input file may have the following arrangement: 1. with sample ID only, 2. With Date/Time only 3. with both Sample ID and Date/Time and lastly 4. without IDs or Date/Time. Names of the species must be unique. As it comes to the unity information, it is not mandatory since those are not included in the uncertainty file. However, information about units may be introduced as a new row in the concentration file.

**Table 3: Input Concentration File**

	PM 2,5	EC	OC	Cl	NO <sub>3</sub>	SO <sub>4</sub>	NH <sub>4</sub>	K	Fe	Zn	Pb
data	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )	(µg/m <sup>3</sup> )
10/08/2002	23.0	0.56	2.78	0.17	4.86	2.52	1.90	0.122	0.056	0.022	0.048
12/08/2002	16.2	0.55	1.96	0.01	0.24	1.60	0.48	0.039	0.058	0.005	0.01
11/09/2002	16.2	2.85	5.73	0.20	0.60	1.60	0.50	0.057	0.134	0.049	0.034
13/09/2002	29.5	1.24	8.91	0.01	2.82	3.72	1.90	0.177	0.129	0.064	0.059
14/09/2002	47.1	1.13	12.33	0.01	7.40	5.45	3.70	0.236	0.160	0.097	0.083
15/09/2002	27.7	1.11	12.80	0.40	2.60	6.50	3.60	0.146	0.103	0.08	0.061
18/09/2002	43.1	1.67	12.31	0.01	17.00	9.10	5.60	0.205	0.198	0.106	0.093
19/09/2002	66.6	2.01	19.47	0.35	13.37	9.03	5.60	0.155	0.129	0.076	0.046
22/09/2002	27.0	0.62	3.54	0.01	5.10	4.60	2.50	0.089	0.102	0.147	0.048
25/09/2002	16.4	2.43	4.21	0.01	3.90	1.00	1.10	0.090	0.119	0.082	0.034
29/09/2002	22.5	1.28	6.19	0.01	2.70	1.70	1.30	0.124	0.125	0.051	0.025
07/12/2002	63.7	2.16	23.16	1.35	7.48	3.85	3.10	0.474	0.228	0.150	0.141
08/12/2002	30.4	1.39	8.40	0.45	4.09	2.58	1.70	0.105	0.053	0.013	0.016
10/12/2002	23.8	0.95	7.58	0.27	6.42	5.32	3.20	0.266	0.128	0.068	0.078
15/12/2002	61.8	1.65	27.74	1.50	11.78	10.35	6.10	0.607	0.234	0.139	0.156

Keep in mind that the blank cells are not permitted, otherwise the user will be asked to examine the data and repeat the process. Likewise, the named species cannot contain commas. Model can recognize negative values, including high values such as -999 and it can proceed, however the model will firstly send the warning message. Still, if the values of sample are not real or if these are missing, the user should revise data outside the program and upload again the data set if necessary.

Sampling and analytical errors can be identified from the sample species uncertainties. In some cases, analytical laboratory can provide an uncertainty assessment for each value. Nonetheless, uncertainties are not always available, thus errors must be estimated by the user.

Uncertainty files are being accepted by EPA PMF 5.0 in only two forms: 1. observation-based and 2. equation-based. The first group is providing estimated information of the uncertainty for each species in a sample. Dimensions should be the same as in the concentration file, but should not include units. The program itself will check the correspondence of the uncertainties file and the concentration file, and user will be notified if some mismatches appear. Nonetheless, program will continue running if encounters some minor mistake, but in a case of mismatching in the number of samples, the program will not allow further evaluation of data. In the uncertainty file, negative or zero values are not considered as relative values and must be excluded from data set. If some values are present with stated values, PMF will show an error message and the user will be asked to remove these values from the file and to reload the uncertainty file.

Uncertainty file that contains species-specific parameters, the software EPA PMF 5.0 processes in order to calculate uncertainties for each sample. This file should contain one row of species with their names. The row that follows should contain species-specific method detection limit (MLD) that is accompanied by the row of uncertainties. As stated before, zero and negative numbers are not permitted for both, detection limit or for the uncertainty values (EPA Positive Matrix Factorization (PMF) 5.0. Fundamentals and User Guide).

## **2.5 Output Data**

The analyst can identify the output file, followed by the choice of the PMF output file types and specify a prefix for the output files. The prefix is always added to the beginning of every file, thus it will be always used as the first part of the output file. Letters or numbers can be used in creation of the prefix, however other characters such as “-” and “\_” are not permitted. In a case that this prefix is not changed during the succeeding run, a warning message will be showed.

After the base runs are finalized, the output files are created. These files contain all necessary information for the on-screen display of the results. The number of the created output files, depends mainly on the type of the output file chosen. Accordingly the type of the output file may be:

- Tab-delimited text (.txt)
- Comma-separated variable (.csv) or
- Excel Workbook (.xls)

If analyst choose Excel Workbook as an output file, two output files will be automatically created by EPA PMF during base runs and will be preserved in the output folder that was selected by analyst: \*\_base.xls and \*\_diagnostics.xls. These files are containing the following information:

- \*\_base.xls - Profiles, Contributions, Residual, Run Comparison
- \*\_diagnostics.xls. - Summary, Input, Base Runs

On the other hand, if the analyst chooses comma-separated variable (.csv), five output files will be created: \*\_diag, \*\_contrib, \*\_profile, \*\_resid and \*\_run\_comparison, where each of the sections expresses the following information:

- \*\_diag has an information of the user inputs and model diagnostic information
- \*\_contrib has the contributions for the each base run which is used to create the contribution graphs on the Profiles/Contributions tab. Run number is governing the sorting of the contributions. Firstly shown are always normalized contributions, and contributions in mass units is following if a total variable is specified.
- \*\_profile has the profiles for every base run which is used to create graphs on the Profiles/Contributions tab. Like in the previous case, profiles are sorted by run number, where profiles in mass units are shown first, followed by profiles in percent of species and concentration fraction of species total if a total mass variable is specified.
- \*\_resid has the residuals, which are regulated and scaled by the uncertainty data for every base run. It is used to create the graphs and tables on the residual analysis screen.
- \*\_run\_comparison has a summary of the species distribution across all PMF runs and for all factors and their comparison to the lowest Q (robust) run.
- \*\_base has the \*\_contrib, \*\_profile, \*\_resid and \*\_run\_comparison in the same Excel Workbook, however in the separated worksheets.

Likewise, if a “.txt” type is selected, the information in the base runs tab is created as separate file and the diagnostics tab information is merged into one file.

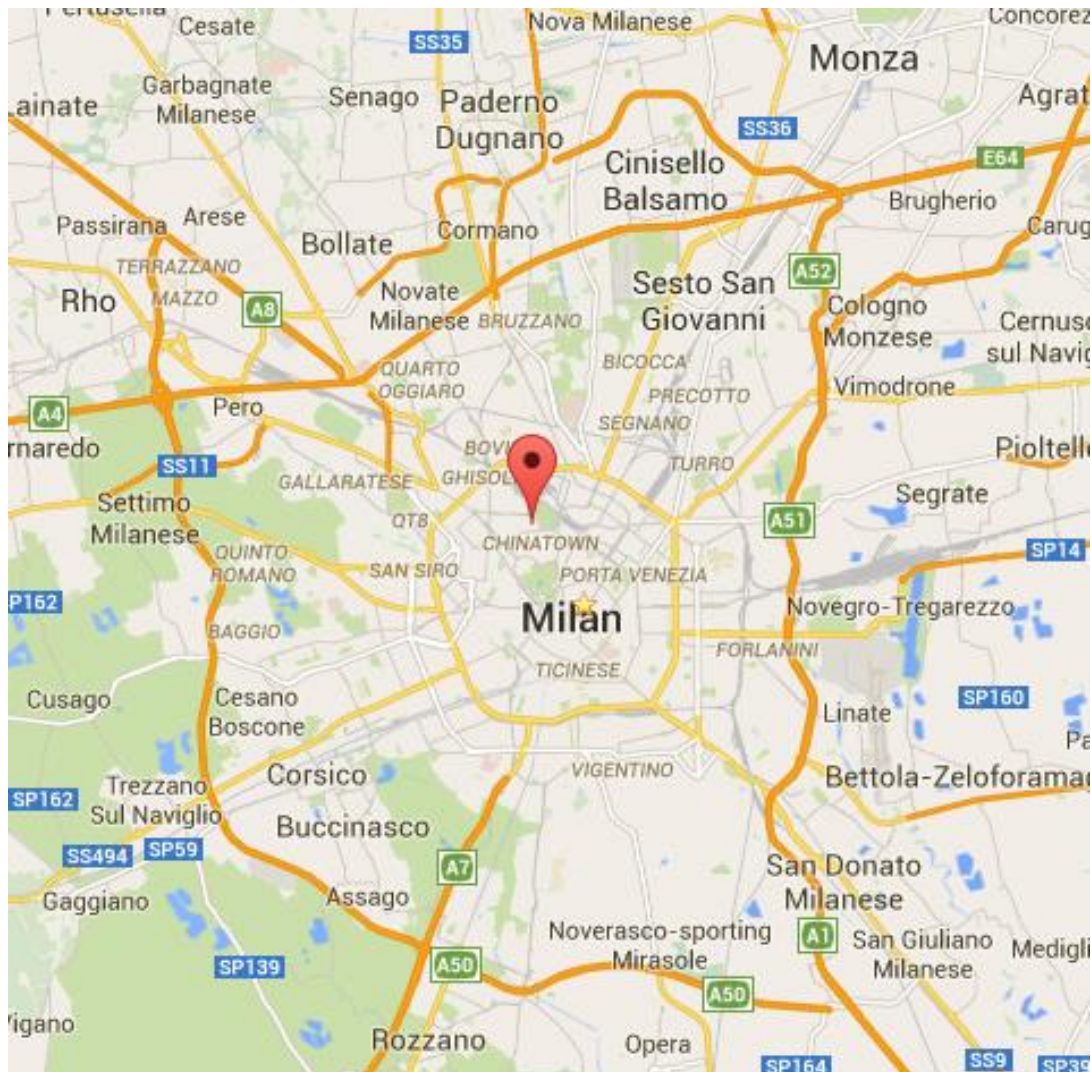
All output files are being saved to the directory previously specified “Output Folder” section in the Data Files screen and with using the prefix defined in the “Output File Prefix” section (EPA Positive Matrix Factorization (PMF) 5.0. Fundamentals and User Guide).

## 2.6 Environmental Data

The data set (Table 3) contains PM<sub>2.5</sub> data set which is derived from sampling campaigns at a monitoring site located in Via Messina in downtown Milan (Figure12) and was conducted by Politecnico di Milano in a period from August 2002 until December 2003.

Sampling campaigns were executed with a high-volume (30m<sup>3</sup>/h flow rate) gravimetric sampler DIGITEL DA-80H, equipped with PM<sub>2.5</sub> cut-off inlet and 150 mm quartz-fiber filters. The location of a sampler was in the walled yard of a residential area, quite far from the major roads. Consequently, this location may be considered as representative of the urban background environment since it was not directly exposed to traffic emissions (Lonati et al. 2008).

Figure 12: Sampling Site



**Table 4: Dataset, Via Messina, Milan**

<b>Sampling site</b>	Milan, Via Messina
<b>Type of environment</b>	Urban area
<b>Sampling period</b>	from 01.04.2002 to 13.12.2003
	<i>(Concentrations have been normalized at 20°C and 101.3kPa)</i>
	<b>PM2.5</b>
<b>N° of samples</b>	162
<b>Warm season</b>	78
<b>Cold season</b>	84
	<b>Carbon (EC and OC)</b>
<b>N° of samples</b>	118
<b>Warm season</b>	64
<b>Cold season</b>	54
	<b>Ions (Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>)</b>
<b>N° of samples</b>	118
<b>Warm season</b>	64
<b>Cold season</b>	54
	<b>Trace elements (Si, K, Fe, Cu, Zn, Pb, Al, Ca, Ti, V, Mn, Ni, Cr, Br)</b>
	- Varying -

The data set contains samples which were analyzed for the PM<sub>2.5</sub> mass (162, 24-h concentration samples), organic carbon, elemental carbon, ionic species Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup> (118, 24-h concentration samples) and trace elements Si, K, Fe, Cu, Zn, Pb, Al, Ca, Ti, V, Mn, Ni, Cr, Br. Among them, organic and elemental carbon were identified by thermal-optical transmission (TOT) method, while ionic components have been ultrasonically extracted from the filter sample by means of high pressure liquid chromatography technique (HPLC). Lastly, trace elements were determined by energy dispersive X-ray fluorescence (XRF) spectrometer (Lonati et al. 2008).

In the original data set some samples did not contain all measurements for some species. Taking into account that PMF model cannot be run if some data is missing, a certain amount of species and samples needed to be excluded from the study; thus, original data set was reduced in order to create a data set that will contain records for all days and species present.

More accurately, from the original data set, mainly trace elements were excluded from the scope of the study, since those were the ones which were missing the most of the measurements. More accurately, Al, Si, Ca, Ti, V, Cr, Mn, Ni, Cu, Br were excluded from the case study, inducing reduction of the number of species from originally 21 up to 11. However, some of the species were still missing some measurement entries, thus exclusion of some sampling dates occurred as well, until the complete dataset was obtained which had no missing measurements. With the reduction procedure, the original data set of 162 concentration samples was reduced to 99 (Table 4). Within the new scope of 99 concentration samples, 48 were measured in the winter season (from October until the end of March), and 51 of them in the summer season (from April until the end of September).

The type of the season is influencing strongly to the outcome of the measurements. The winter season lasts from 01.10. until the 31.03., while the Summer season lasts from 01.04. until the 31.09. During the winter season, despite the traffic, the heating on various fossil fuels is present. Combined together, these two aspects are causing pick values in the concentration levels of observed elements. While, on the other hand during the summer season, heating is not present in this region. Thus, causing the lowest values in the records obtained.

**Table 5: Dataset after Reduction**

<b>Sampling site</b>	Milan, Via Messina
<b>Type of environment</b>	Urban area
<b>Sampling period</b>	from 10.08.2002 to 13.12.2003
	<i>(Concentrations have been normalized at 20°C and 101.3kPa)</i>
	<b>PM2.5, Carbon (EC and OC), Ions (Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>)</b>
<b>N° of samples</b>	99
<b>Warm season</b>	50
<b>Cold season</b>	49
	<b>Trace elements (K, Fe, Zn, Pb)</b>
	- Varying -

Preliminary statistical analysis, i.e. mean concentration, standard deviation as well as minimum and maximum values of all data set after reduction is presented in the Table 5.

**Table 6: Principal Statistical Parameters of Samples**

Samples	max	min	mean	st.dev
PM 2,5	133.42	12.18	43.24	27.10
EC	4.97	0.50	1.49	0.71
OC	35.78	1.96	10.16	7.47
Cl	2.52	0.01	0.36	0.55
NO <sub>3</sub>	71.62	0.24	9.56	11.77
SO <sub>4</sub>	15.31	0.46	4.79	2.61
NH <sub>4</sub>	16.16	0.08	3.13	2.74
K	0.94	0.01	0.10	0.14
Fe	0.79	0.05	0.24	0.14
Zn	0.29	0.001	0.06	0.05
Pb	0.35	0.01	0.05	0.05

The values of PM<sub>2.5</sub> are varying from 12.18 µg/m<sup>3</sup> measured during the warm season up to 133,4 µg/m<sup>3</sup> obtained in the winter season. Corresponding differences in picks are primary due to a seasonal variation, i.e. presence/absence of emissions from domestic heating in cold and warm season, respectively. Secondly, differences are caused by variable meteorological conditions where for instance lower mixing height changes (Hueglin et al. 2005). The lowest values (<20 µg/m<sup>3</sup>) were sampled in the end of May and begging of July, while on the contrary the highest values, which were exceeding 110 µg/m<sup>3</sup>, were measures in the end of February and in the beginning of March. Despite that, the mean PM<sub>2.5</sub> mass concentration is 43.24 µg/m<sup>3</sup>.

The biggest standard deviations are being recorder within PM<sub>2.5</sub> and NO<sub>3</sub>, 27.1 µg/m<sup>3</sup> and 11.8 µg/m<sup>3</sup> respectively, while the smallest values are recorded within Pb, Zn, Fe and K, where the standard deviation is less than 0.2.

The data set contains many measurement levels in many categories; however, the analytical findings do not allow determination of the total PM mass due to several reasons:

- Some elements, for instance vanadium, cadmium, etc. have not been determined,
- The residual humidity of the PM sample is not considered,
- X-ray inquiries refers only to elements, while the elements in the particulate matter are in the form of oxides,
- Organic compounds are present as organic matter and not only as organic carbon (Almeida et al. 2005)



## 3 Results and Discussion

### 3.1 PMF Objective

The goal of Positive Matrix Factorization (PMF) is to identify the number of factors  $p$ , the species profile  $f$  of each source, and the amount of mass  $g$  contributed by each factor to each individual sample.

It is very important for the data analyst to know what types of sources are present in the study area. However, in some particular cases where there is a good emission inventory it may happen that a source cannot be identified. Additionally, atmospheric processes may influence on various factors such as summer and winter secondary sulphate, or in generating sufficiently collinear sources that an irresolvable mixture of source profile is generated. Considering that, profiles should be interpreted with both knowledge of the study area and a background in atmospheric science (Belis et al. 2014).

Once the factors are identified, it is necessary to compare obtained factor profiles with those reported in previously published PMF studies or to look at temporal patterns for expected behaviors, e.g. the largest contributions of a source believed to be residential wood burning, should likely occur during winter months etc. Additionally, plots of contribution over time can be inspected in order to look for possible seasonal or yearly oscillations of the source contribution (Farao et al. 2013)

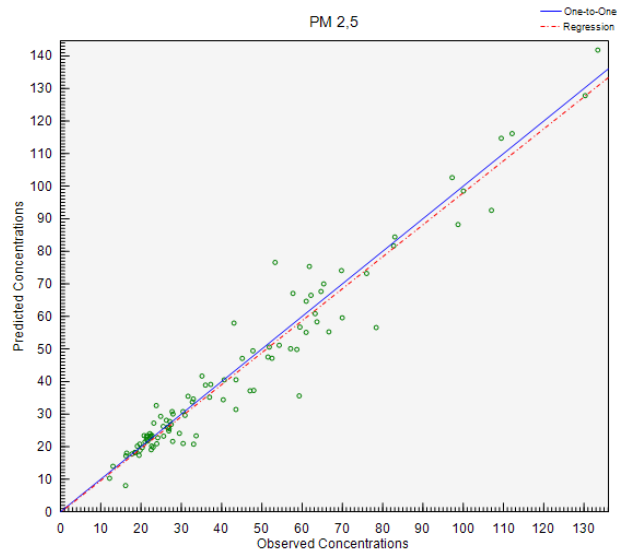
#### 3.3.1 PMF Application

As mentioned in one of the previous paragraphs (2.6 Environmental data) the original dataset have been reduced in order to enable run of the model, following the special criteria of the PMF mode. Firstly the number of species was reduced from 21 up to 11, excluding mainly the trace elements. However, since some species were still missing measurement entries, exclusion of the sampling dates needed to take place as well. Thus, from starting number of 21 species and 162 concentration samples, the dataset ran in the model contained 11 species and 99 concentration samples.

This resulting dataset was input in the model, including the uncertainty dataset of the measurement technique. Uncertainty data set is used to assign to each species and sample additional uncertainty dataset. This uncertainty dataset is assigned to a value of 5%. The dimension of the uncertainty dataset and the dataset of the measurements is the same. The model was iterated 100 times for the development of the most accurate solution, and obtained results are showed and discussed in the subsequent sections.

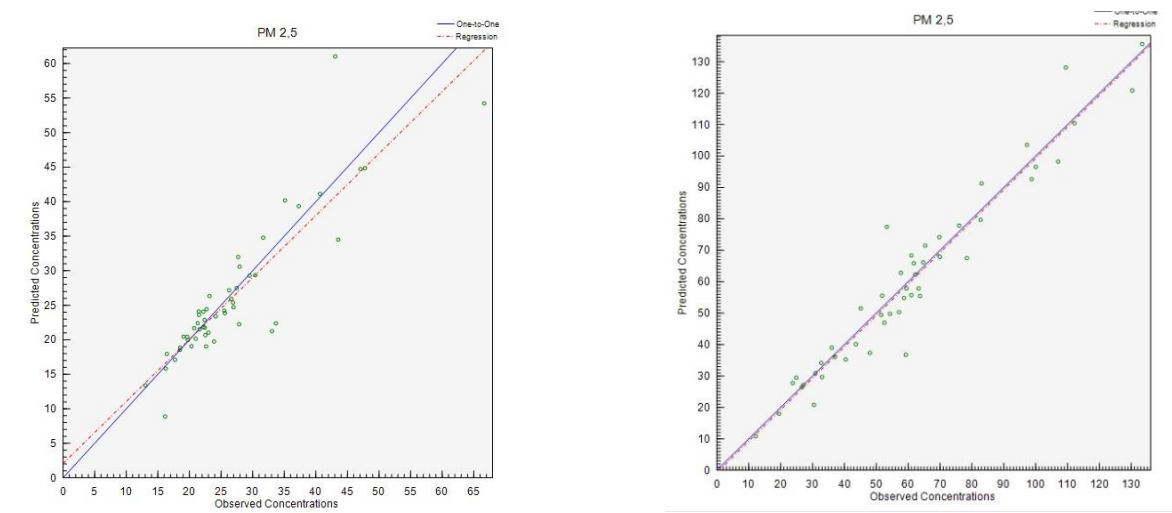
Considering the investigation of dataset, the correlation between measured and reconstructed mass is showed on the Figure 13. The scatter plot shows strong correlation between masses of the  $PM_{2.5}$ .

**Figure 13: Reconstructed  $PM_{2.5}$  Concentrations vs. Measured  $PM_{2.5}$**



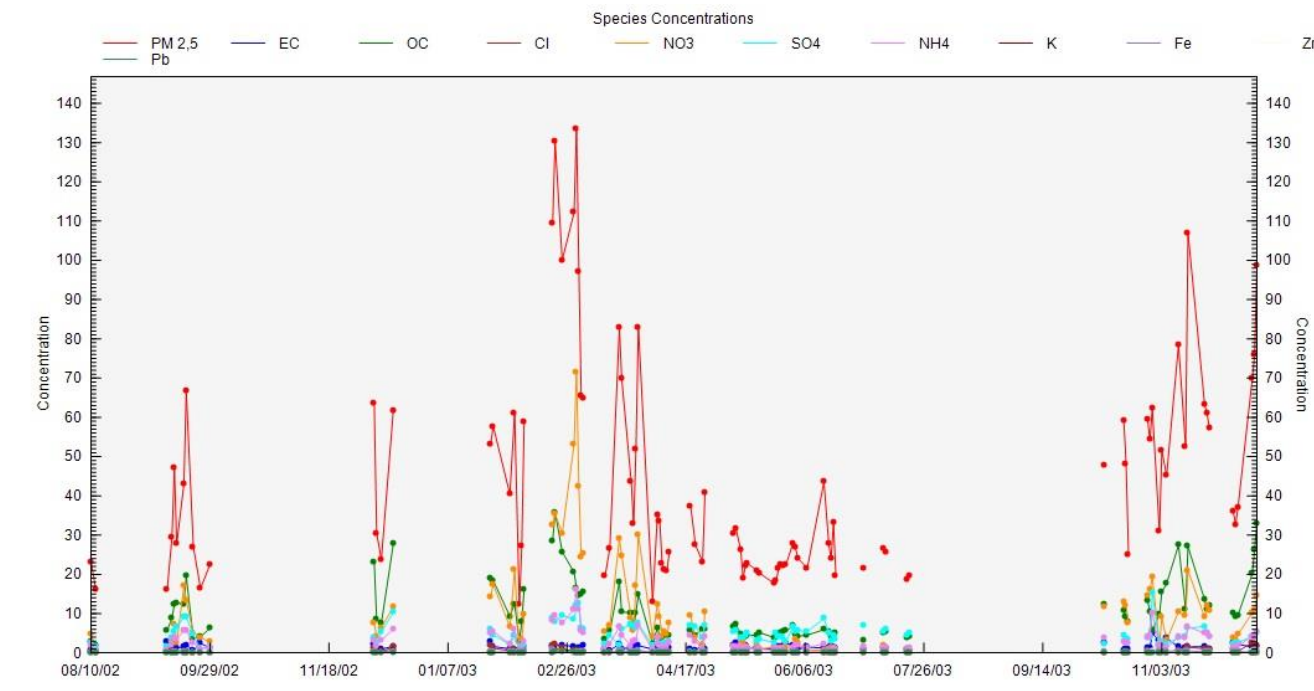
All the rows in the dataset have been considered for the analysis; however it was possible to divide the dataset in two seasonal subsets with approximately 50 observations for each season, in order to achieve the additional analysis. During the processing application of the warm and cold season respectively, the scatter plot of the correlation between measured and reconstructed values of  $PM_{2.5}$  was showing strong trend as well. Thus, Figure 14 shows reconstructed  $PM_{2.5}$  concentration versus measured  $PM_{2.5}$  for both, warm and cold season respectively.

**Figure 14: Reconstructed  $PM_{2.5}$  Concentrations vs. Measured  $PM_{2.5}$  – Warm and Cold Season, respectively**



Concentration time series as the species contribution is showed on the following Figure15. It can be seen that the biggest contributions are from the total PM<sub>2.5</sub> reconstructed mass, NO<sub>3</sub> and OC, where the rest of the observed species are contributing on a smaller scale.

**Figure 8: Concentration Time Series ( $\mu\text{g}/\text{m}^3$ )**



### 3.3.2 PMF Profiles

The dataset has been studied, after being processed and run by the PMF model, following the instructions described in the protocol of the model. The obtained results are shown in the section that follows, accompanied with the detailed interpretation of the findings.

Source profiles, “factors” have been identified with PMF model. Considering the whole dataset, six factors were identified. On the contrary, with respect to the warm and cold season subsets, only 4 factors were identified. All the tables with the profiles of the sources as well with their fingerprints i.e. distribution of sources for PM<sub>2.5</sub>, and factor contributions, are represented.

Figure 16: Factor Contribution

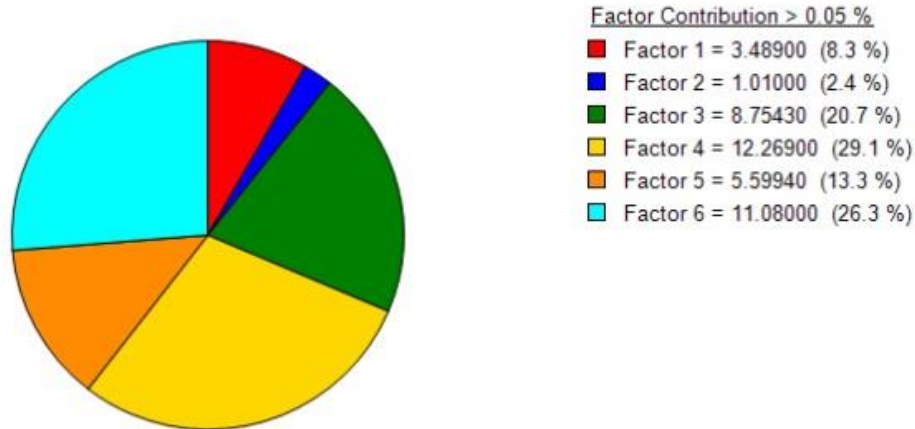


Figure 17: Factor Fingerprint

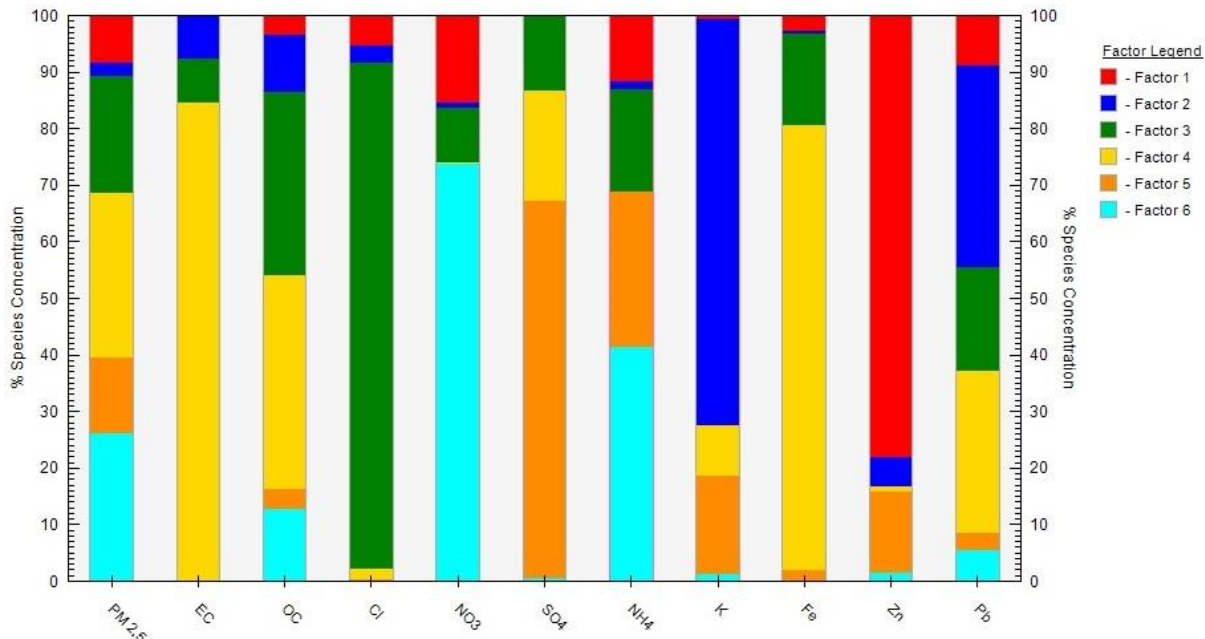
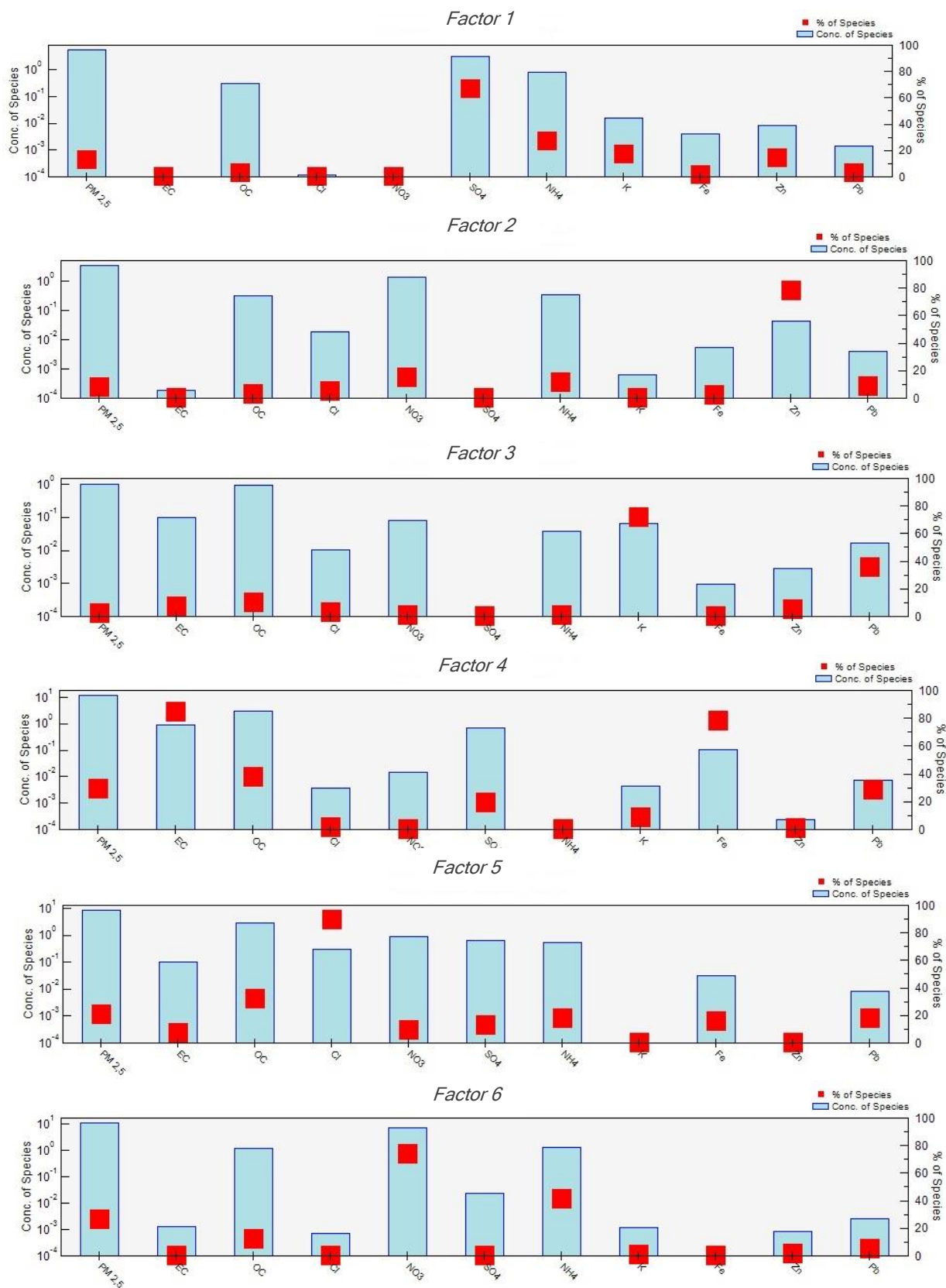


Figure 16 reports factors contribution (i.e.: the share of the total PM<sub>2.5</sub> associated to each of the factors) obtained for the whole dataset. Clearly, factor 4 contributes the most to the total mass of the PM<sub>2.5</sub> with almost 30% of share followed by factor 6 and factor 3 with 26,3% and 20,7% respectively. The rest of the factors are contributing on a much smaller scale, thus with less than 24%, if considering the contribution of all three factors. Accordingly, on the Figure 17, the fingerprint of all factors is showed. By observing this graph, an analyst may have deeper insight into the scope of the contributions, forasmuch both, the share and the species division of the obtained factors. Lastly, in the Table 7, all identified source profiles are showed.

**Table 7: Factor Profiles ( $\mu\text{g}/\text{m}^3$ )**



Likewise, the sources represented by the profiles and main average contributors to their presence in PM<sub>2.5</sub> samples are identified as:

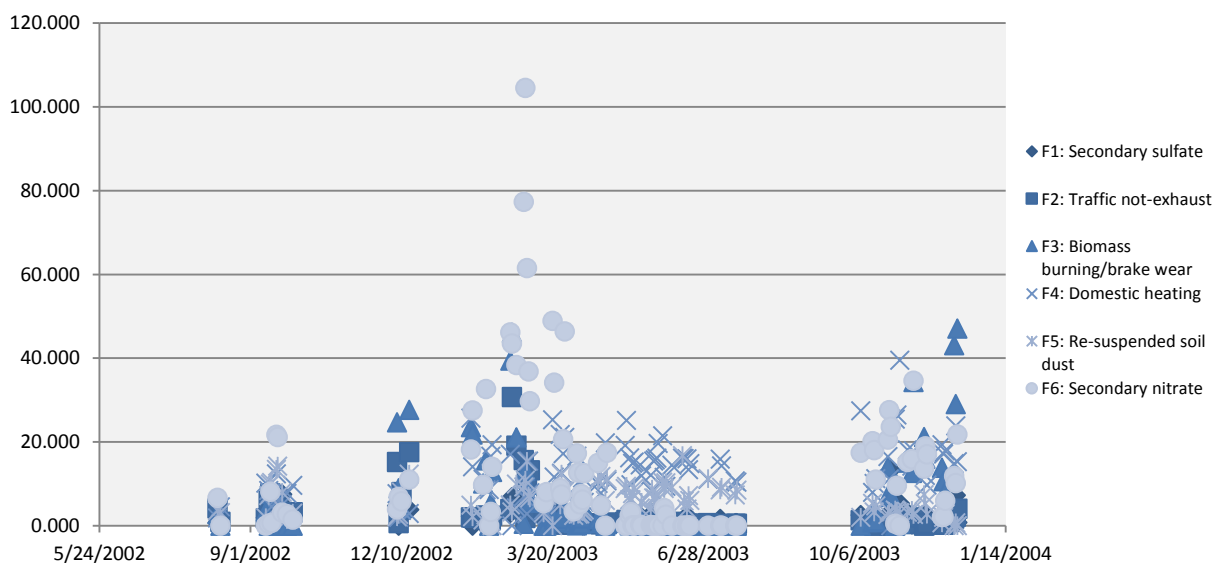
- Factor 1: SO<sub>4</sub> (90%), NH<sub>4</sub> (80%), OC (70%)
- Factor 2: NO<sub>3</sub> (90%), NH<sub>4</sub> (75%), OC (70%), Zn (58%)
- Factor 3: OC (95%), EC (70%), NO<sub>3</sub> (70%), K (70%), NH<sub>4</sub> (60%)
- Factor 4: OC (85%), NO<sub>3</sub> (75%), SO<sub>4</sub> (73%), NH<sub>4</sub> (70%), Cl (60%), EC (55%)
- Factor 5: OC (88%), EC (80%), SO<sub>4</sub> (70%), Fe (58%)
- Factor 6: NO<sub>3</sub> (90%), NH<sub>4</sub> (80%), OC (80%), SO<sub>4</sub> (52%)

By investigating the most recent studies on the apportionment of the PM<sub>2.5</sub> sources, it may be concluded that:

- Factor 1 can be identified as a “Secondary sulfate” source with a small contribution of about 8% to the total mass of PM<sub>2.5</sub>
- Factor 2 is identified as “Traffic non-exhaust” source, with high values of Zn share, mainly due to tire wear or brake material
- Factor 3 stands for “Biomass burning/brake wear”, considering the high contributions of EC and K, and it stands for the third biggest contribution to the mass of PM<sub>2.5</sub>
- Factor 4 is interpreted as “Domestic heating” source, characterized by high relative contribution of SO<sub>4</sub> and likely related to industrial source contribution, representing one of the main contributors to the PM<sub>2.5</sub> mass concentration
- Factor 5 represents the “Re-suspended soil dust”, mainly due to high contributions of EC and Fe
- Factor 6 may be identified “Secondary nitrate”, accounting for the second largest contribution to the PM<sub>2.5</sub> mass concentrations (Farao et al. 2013, Larsen et al. 2012, Piazzalunga et al. 2011, Scotto et al. 2013, Vecchi et al. 2008, Viana et al. 2008)

On the Figure 18, source concentration with respect to time series is depicted. This kind of plot gives a broader overview of the source concentration during time. It may be expected that domestic heating, contributes on a smaller scale in the warm season, thus we may observe this pattern on the plot. Similar trend is identified for the burning of biomass/break wear and secondary nitrate, which contribute vastly in the cold season as heating is immensely used. Traffic not-exhaust source shows increased levels of concentration in the winter time, while re-suspended soil dust and secondary sulfate show relatively uniform trends.

**Figure 18: Source Concentration (temporal series) ( $\mu\text{g}/\text{m}^3$ )**



Data for the warm season subset only, was run with PMF, and the obtained results are reported in the Figures 19 and 20 and in the Table 8.

**Figure 19: Factor Contribution Warm Season**

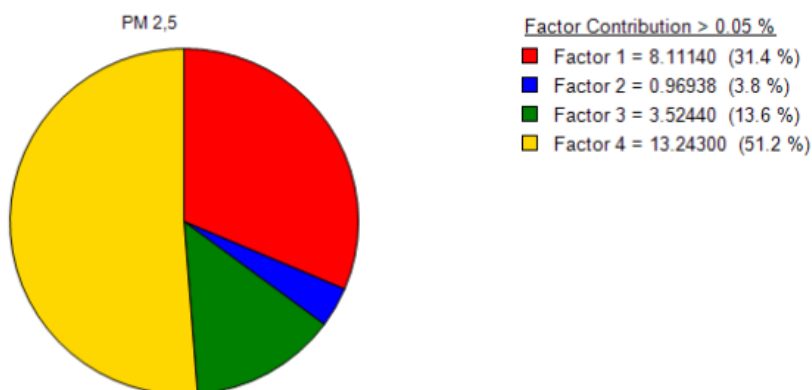
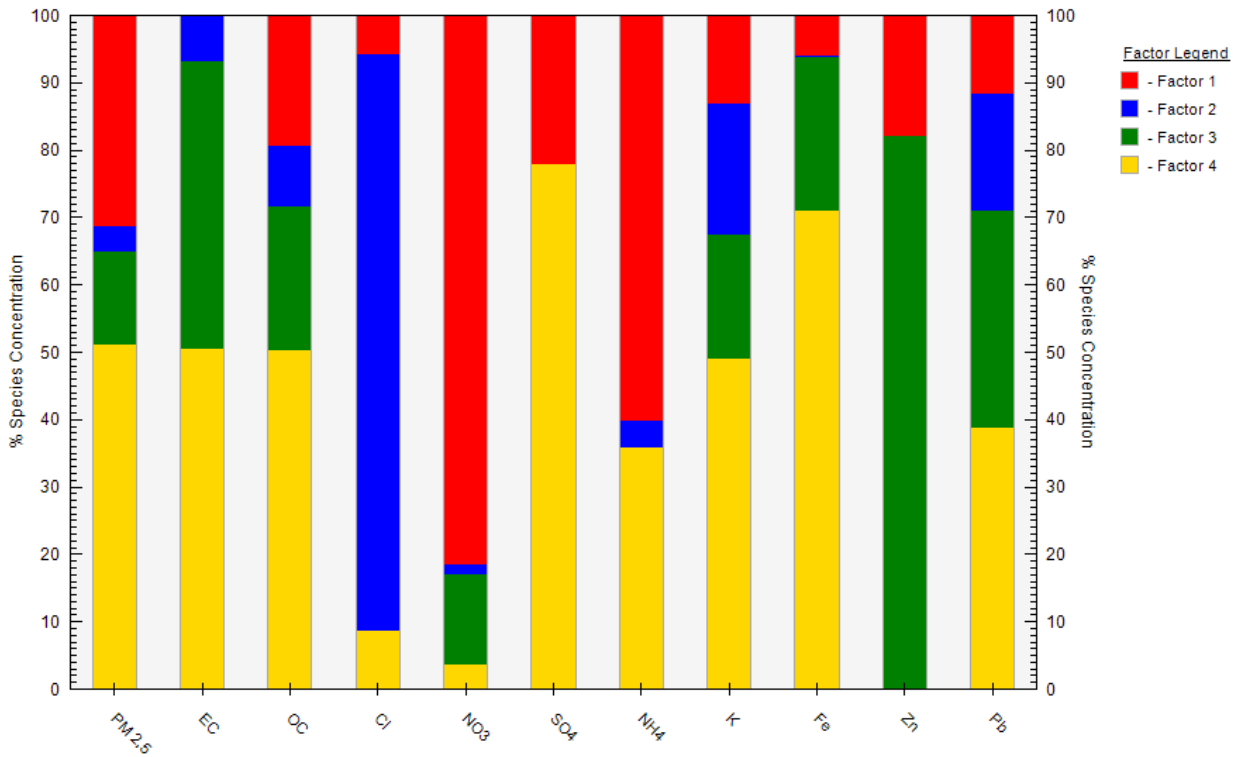


Figure 20: Fingerprint Warm Season



As mentioned before, for the seasonal subsets only four factors are identified. Observing the accessed factors, the one that contributes vastly (51%) to the total mass of the PM<sub>2.5</sub> is factor 4, while the remaining three factors are contributing with a smaller share. More accurately, factor 1 with 31.4%, factor 3 with 13.6% and lastly factor 2 with 3.8% (Figure 19). On the Figure 20, the fingerprint is showed with more detailed information, as for the total dataset in the previous paragraph.



**Table 8: Factor Profiles Warm Season ( $\mu\text{g}/\text{m}^3$ )**

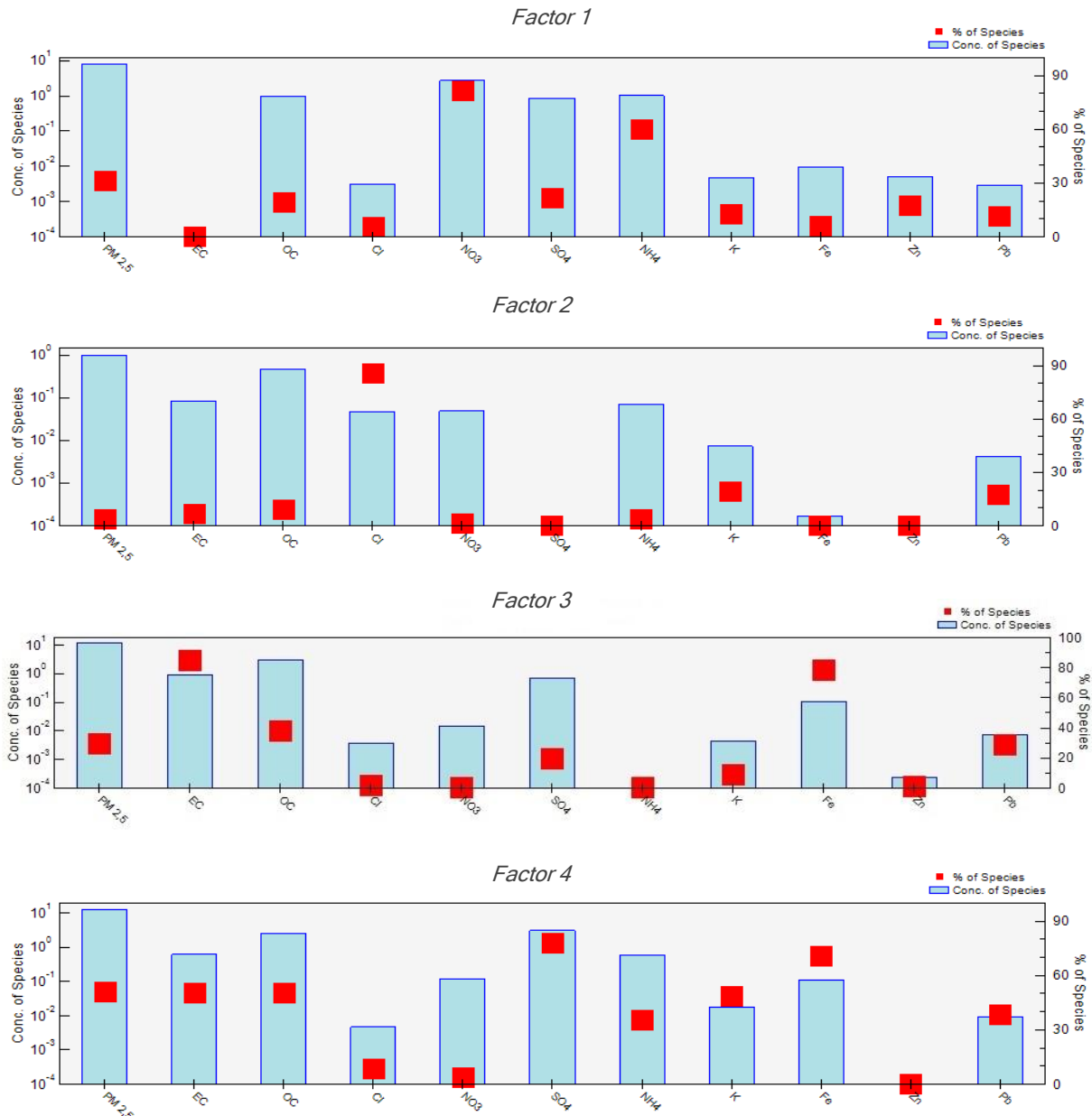


Table 8, depicts the factors identified by the run of the warm season subset.

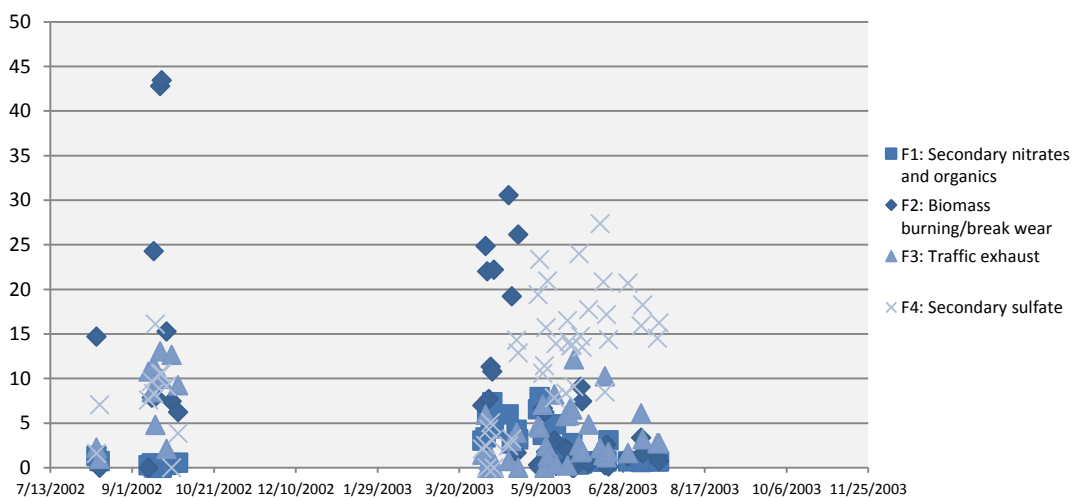
Similarly, the sources represented by the profiles and main average contributors to their presence in PM<sub>2.5</sub> samples are identified as:

- Factor 1: NO<sub>3</sub> (85%), OC (80%), NH<sub>4</sub> (80%), SO<sub>4</sub> (80%),
- Factor 2: OC (90%), EC (72%), NH<sub>4</sub> (70%), NO<sub>3</sub> (70%), Cl (70%)
- Factor 3: OC (85%), EC (75%), SO<sub>4</sub> (75%), Fe (55%)
- Factor 4: SO<sub>4</sub> (85%), OC (82%), EC (70%), NH<sub>4</sub> (70%), Fe (55%), NO<sub>3</sub> (55%),

Using the same approach as for the whole dataset profiles identification, the following sources are recognized:

- Factor 1 may be identified as “Secondary nitrates and organics”, accounting for the second largest contribution to the PM<sub>2.5</sub> mass concentrations
  - Factor 2 assumed to be “Biomass burning/brake wear” source, generally accounted for the break material and tire wear and it stands for the smallest contributor to the mass of PM<sub>2.5</sub>
  - Factor 3 may be recognized as the “Traffic exhaust”, being one of the significant contribution to the total mass of PM<sub>2.5</sub>
  - Factor 4 is interpreted as the “Secondary sulfate”, representing the greatest contributors to the PM<sub>2.5</sub> mass concentration
- (Farao et al. 2013, Larsen et al. 2012, Piazzalunga et al. 2011, Scotto et al. 2013, Vecchi et al. 2008, Viana et al. 2008)

**Figure 21: Source Concentration Warm Season (temporal series) (µg/m<sup>3</sup>)**



On the Figure 21 it may be identified the temporal series of the source concentration for the warm season subset. By observing the figure, two peaks may be identified in September, where mainly biomass burning/brake wear and traffic exhaust are showing high concentration trends. It may be assumed that biomass burning/brake wear starts to contribute greater as a biomass burning, since the cold season is approaching. Likewise, the season of holidays is over, thus the traffic exhaust starts to contribute significantly. On the other hand, secondary nitrate has relatively uniform trend, while secondary sulfate has the highest concentration levels in the spring time. Additionally, it should be seen that the model doesn't recognize a domestic heating source, since warm season dataset is being considered.

On the other hand, by observing the cold season subset, the subsequent results are identified:

Figure 22: Factor Contribution Cold Season

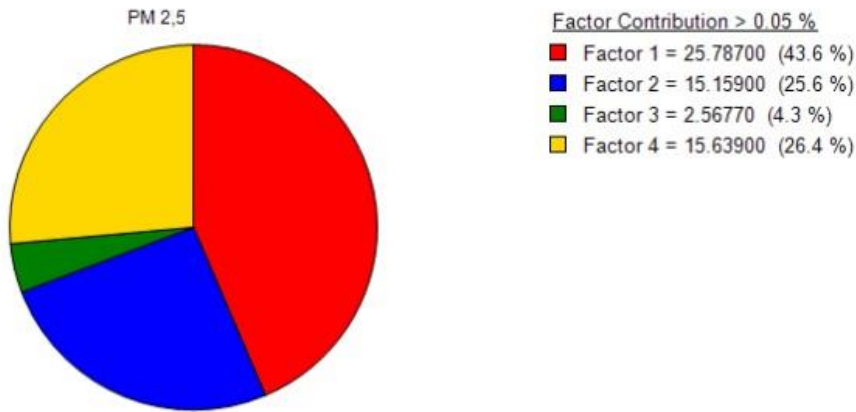
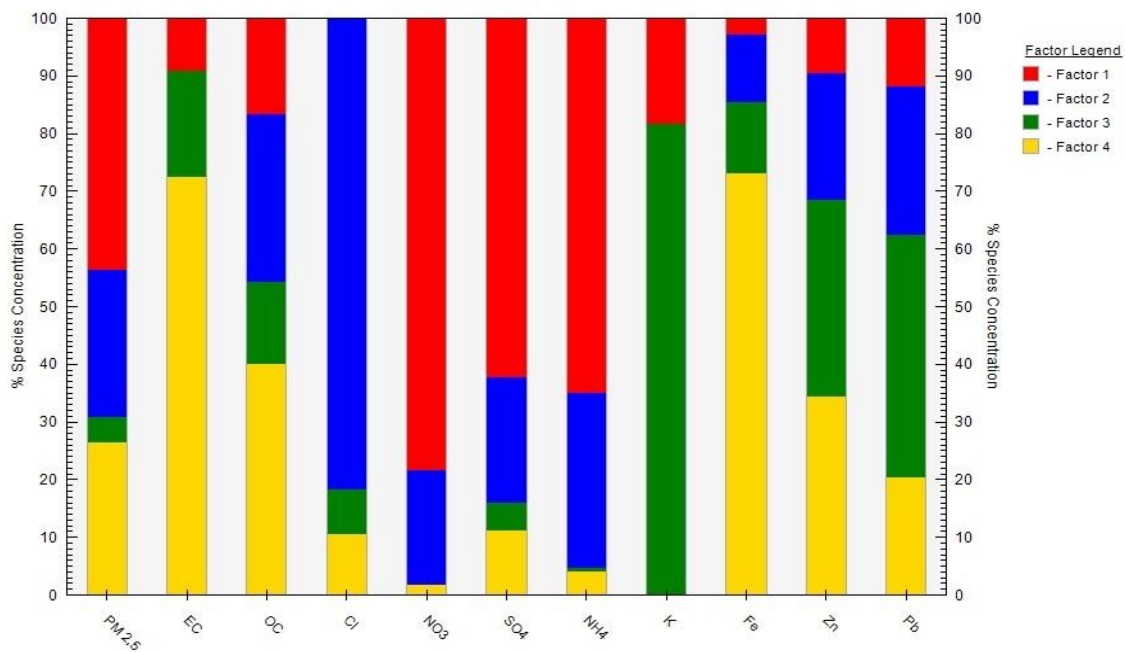
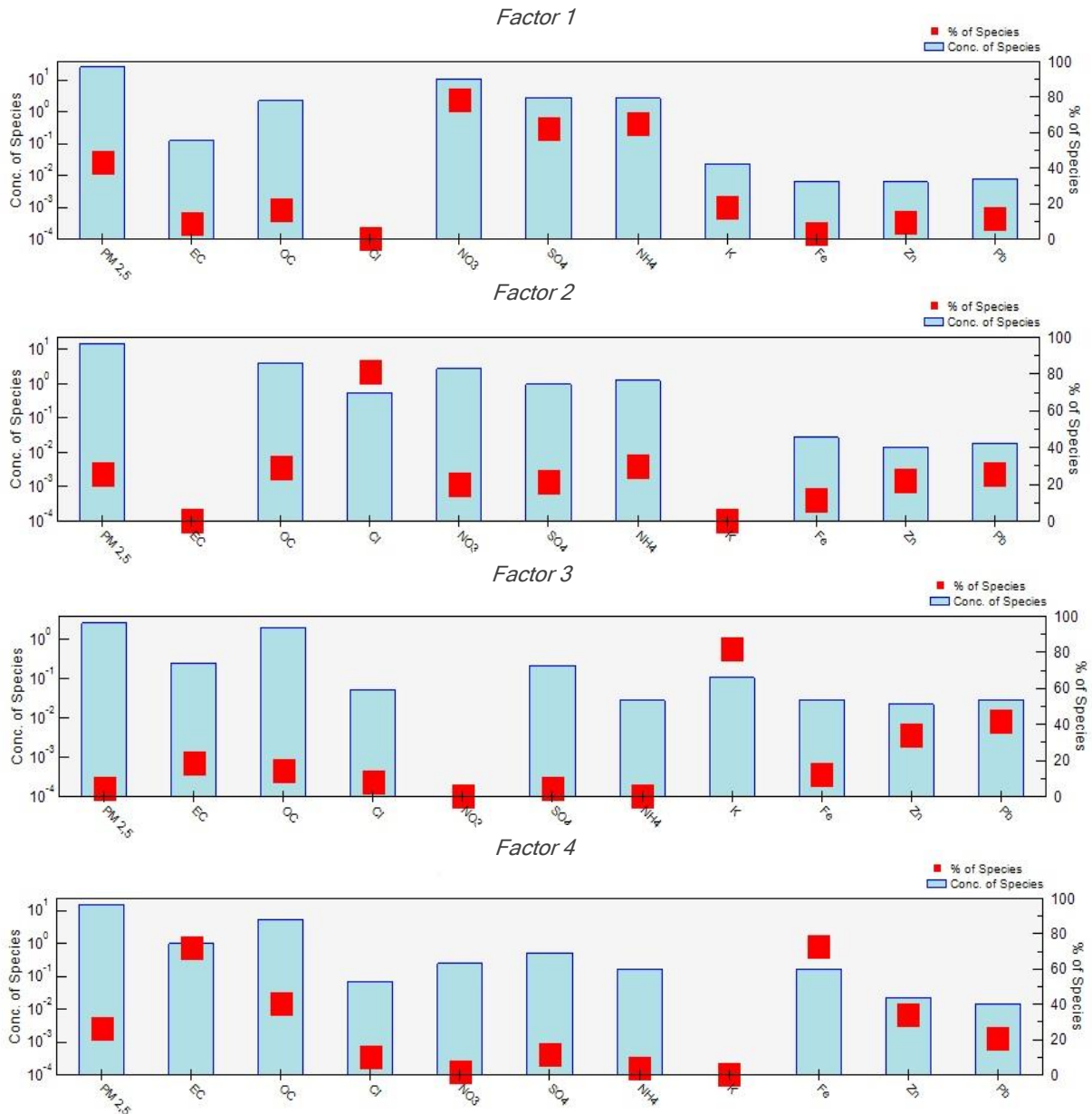


Figure 23: Fingerprint Cold Season



The factor that dominates the most in the share of contribution to the total mass of PM<sub>2.5</sub> is the factor 1 with 43.6%, as showed in the Figure 22. The second biggest influencer is shared between factor 2 and factor 4, with 25.6% and 26.4% respectively. The factor with the lowest weight towards the contribution is the factor 3 with only 4.3%. Lastly, on the Figure 23, the fingerprint of all factors with all observed species is presented.

**Table 9: Factor Profiles Cold Season ( $\mu\text{g}/\text{m}^3$ )**



In the Table 9 obtained profiles for the cold season are depicted.

Additionally, the sources represented by the profiles and main average contributors to their presence in PM<sub>2.5</sub> samples are identified as:

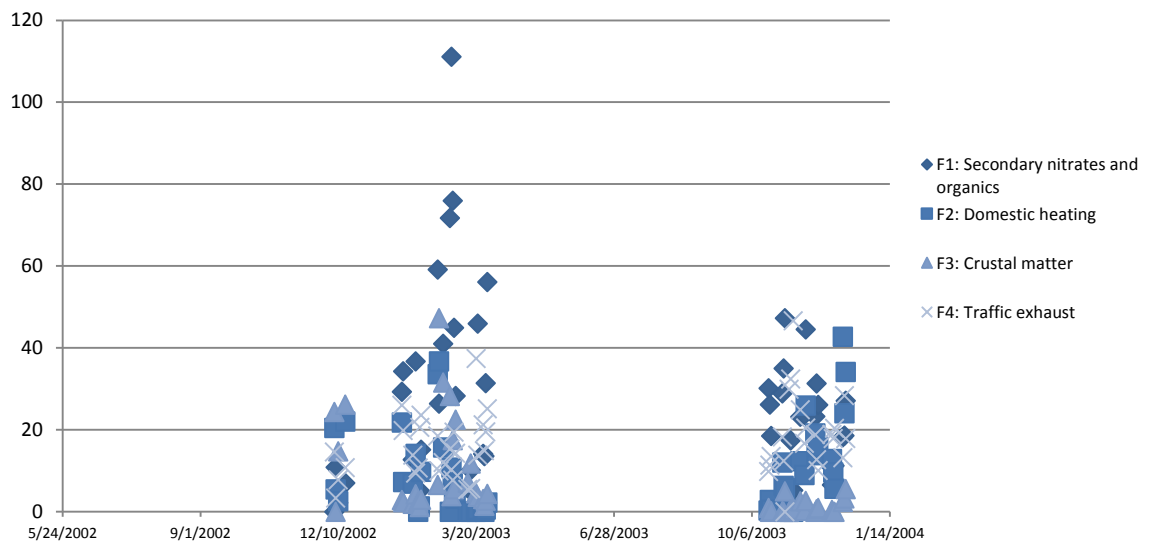
- Factor 1: SO<sub>4</sub> (80%), NH<sub>4</sub> (80%), OC (77%),
- Factor 2: OC (85%), NO<sub>3</sub> (85%), NH<sub>4</sub> (78%), SO<sub>4</sub> (75%), Cl (70%)
- Factor 3: OC (90%), EC (75%), SO<sub>4</sub> (75%), K (70%), Cl (60%), NH<sub>4</sub> (55%), K (55%)
- Factor 4: OC (90%), EC (75%), SO<sub>4</sub> (70%), NO<sub>3</sub> (65%), NH<sub>4</sub> (60%), Fe (60%),

By using the same approach as for the total dataset profiles identification, the following sources are recognized:

- Factor 1 is identified as a “Secondary nitrates and organics” source with greatest contribution of about 43% to the total mass of PM<sub>2.5</sub>
- Factor 2 assumed to be “Domestic heating”, that stands for the third biggest contribution to the mass of PM<sub>2.5</sub>
- Factor 3 may be recognized as “Crustal matter<sup>5</sup>” source, with high values of Zn share, mainly due to tire wear or brake material
- Factor 4 is interpreted as the “Re-suspended soil dust”, mainly due to high contributions of EC and Fe, representing the second main contributors to the PM<sub>2.5</sub> mass concentration

(Farao et al. 2013, Larsen et al. 2012, Piazzalunga et al. 2011, Scotto et al. 2013, Vecchi et al. 2008, Viana et al. 2008)

Figure 24: Source Concentration Cold Season (temporal series) (µg/m<sup>3</sup>)



Temporal series for source concentrations, considering only the cold season is depicted on the Figure 24. As expected, domestic heating is identified again, since in the warm season its contribution was rather negligible, where other sources were influencing it on a much larger scale. The greatest pick is identified for February and March, as being one of the coldest months where heating reaches its maximum. For the same reason, high levels of secondary nitrates and organics are obtained. Lastly, traffic exhaust show relatively significant and uniform trend.

<sup>5</sup> Crustal Matter is a fugitive dust originating from the Earth’s crust

While pairing resulting profiles, all above identified source profiles are influenced by subjective opinion of the analyst. The obtained mathematical results from the model run were compared and matched with previously published source profile's articles. The assigned nomination of the source profiles was taken from the published articles, keeping in that way the consistency of the profiles. Lastly, due to subjective stance of the analyst and some complex or unclear profiles, some minor omissions may occur in the interpretation of the reported source profile's nomination.

For more detailed investigation, concerning the statistical background of the results obtained, box whisker plot<sup>6</sup> PM<sub>2.5</sub> mass contribution is showed in the Figure 25. More particularly, the plot depicts sets of numerical data through their quartiles, where variability outside the upper and lower quartiles may be observed, i.e. they represent distribution of the estimated daily contribution of the identified sources.

By observing the Figure 25, it may be concluded that Factor 4 contributes the most to the total PM<sub>2.5</sub> mass concentration, since its median is having the highest value of 31%. In addition, Factor 4 is followed by Factor 5 that has a median value that corresponds to 29%. It should be noted that Factor 5 is distributed wider than the Factor 4 which is mainly concentrated around its median value. Factor 6 have very similar distribution as Factor 5 has, while Factor 1 shows the highest distribution of its results among all profiles. Moreover, contributions towards the total mass of PM<sub>2.5</sub> are for the Factor 1 (15%) and for the Factor 6 (14%). As the smallest contributors with a moderate distribution around its median values are Factor 3 (5%) and Factor 2 (4%).

Lastly, Figure 26 shows the factor contribution dependency on PM<sub>2.5</sub> concentration values. Traffic non-exhaust, domestic heating and secondary nitrate are not influenced by the concentration values of PM<sub>2.5</sub>, hence these factors are high degree insensitive to its levels. On the other hand, starting from biomass burning/brake wear, across the secondary sulfate, up to the re-suspended soil dust, are sources positively dependent on PM<sub>2.5</sub> concentration, respectively.

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<sup>6</sup> Box Whisker Plot: symbolizes conducive way of graphically rendering sets of numerical data through their quartiles. It may also include lines spreading vertically from the boxes (whiskers) that demonstrate variability outside the upper and lower quartiles.

Figure 25: Box plot for the PM2.5 Mass Contribution

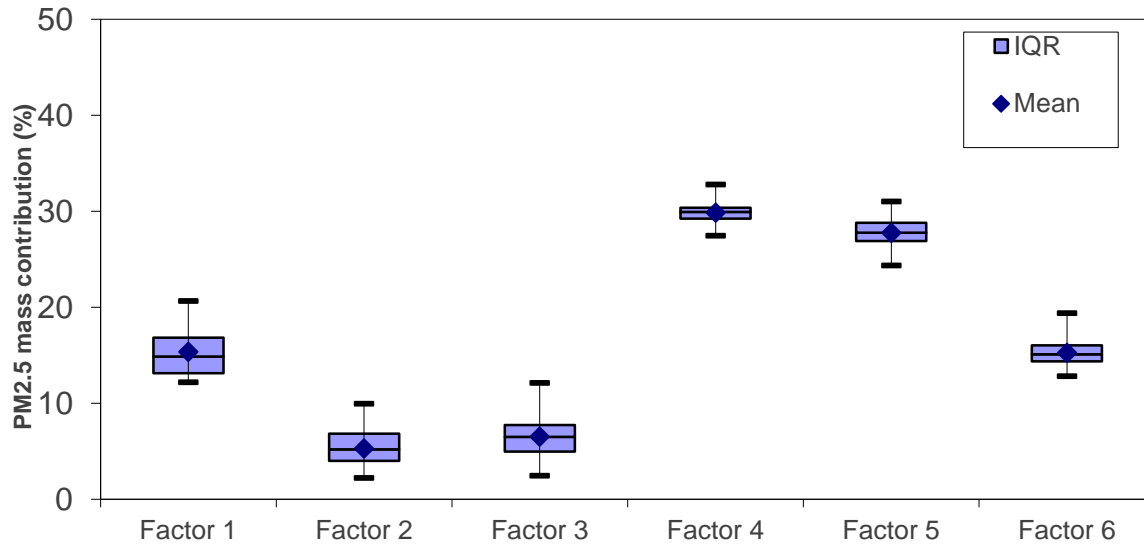
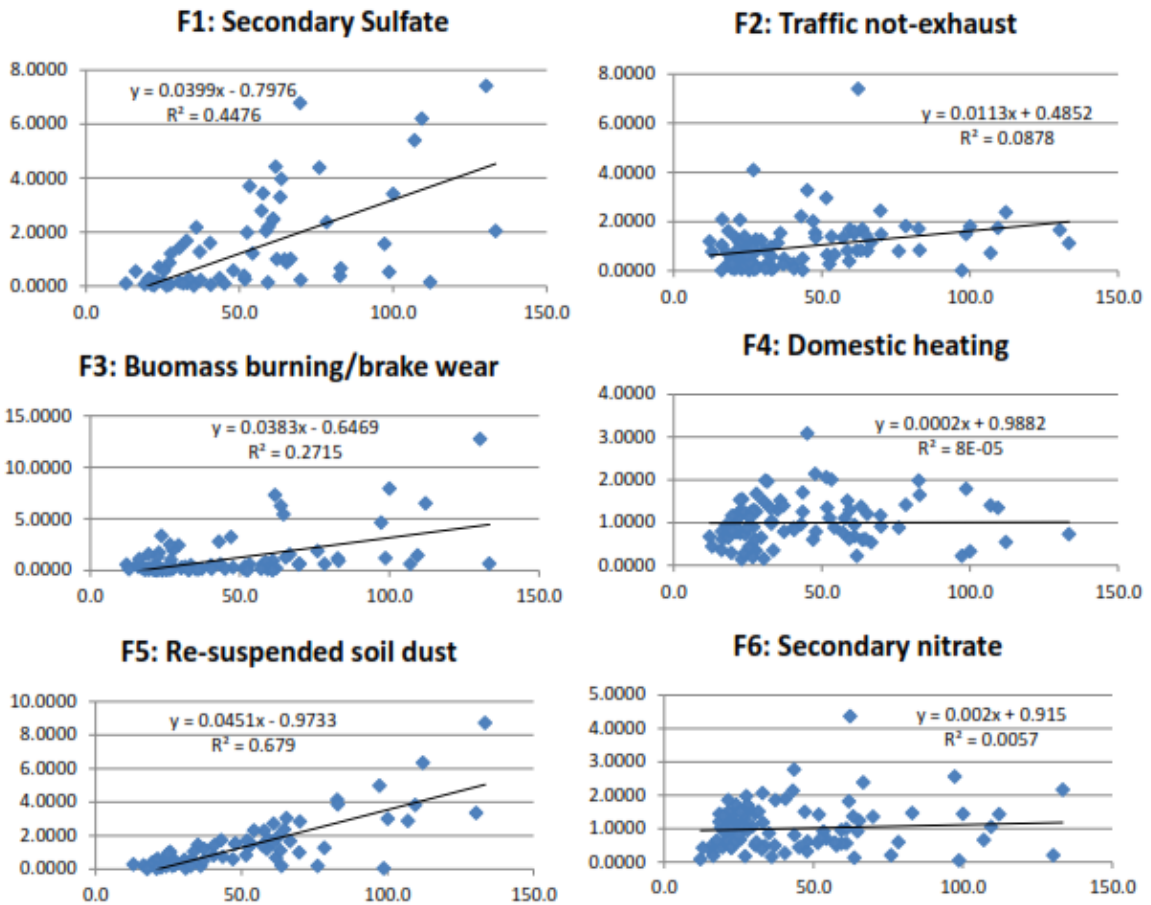


Figure 26: Daily to Mean Ratios for all Sources in the Annual Dataset



### 3.3.3 Comparison to Other Receptor Models

Besides Positive Matrix Factorisation (PMF) model, UNMIX and Chemical Mass Balance (CMB) models have used the same dataset as well, in order to study the source profiles from the PM<sub>2.5</sub> in the Lombardy region. These case studies were performed by students from Politecnico di Milano. More particularly, this dataset was run and investigated with CMB model (A. V. Iannucci and M. Campagnoli) and secondly with UNMIX model (G. Sporchia and U. S. Nobile) as part of the Master thesis. Since, pure research has been performed on the source profiles for this particular area, obtained results from all three models were compared and conclusion is reported in the following paragraph.

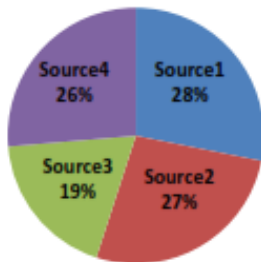
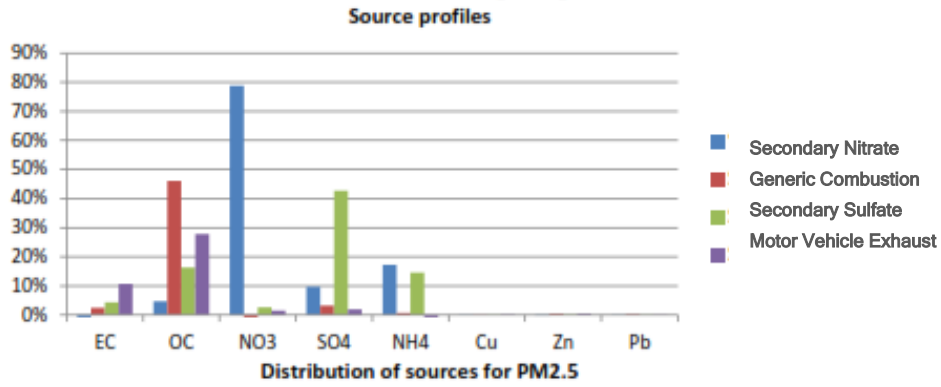
After the run of the annual dataset by PMF and UNMIX models, PMF model recognized the most of source profiles, thus six, while UNMIX model highlighted 4 profiles. CMB has identified 4 source profiles, but only with respect to seasonal subset; hence it will be commented afterwards. In the both models output, ones relevant to annual dataset, same source profiles were identified, secondary sulfate and secondary nitrate. Since, PMF highlighted more profiles, it is expected that the model output did not identify same general source profiles, instead, other similar ones more specific. Therefore, UNMIX model has recognized as the two last profiles, motor vehicle exhaust and generic combustion. The remaining profiles from PMF run were identified as domestic heating, biomass burning/break wear, re-suspended soil dust and lastly, traffic non-exhaust. However, the number of source profiles is not identical and comparison among their contribution share in total PM<sub>2.5</sub> concentration cannot be accomplished, but the biggest contributors may be recognized. As a result, the most contributing profile in the UNMIX model run output is shared among secondary nitrate and generic combustion source, while on the contrary, the ones contributing the most in the PMF model run output were domestic heating and secondary nitrate. By observing these findings, it may be said that the models were quite similar in their results, since combustion source and domestic heating can be considered as almost the same one.

On the figures 27, 28 and 29, the summary of UNMIX and PMF sources, and their average contribution to PM<sub>2.5</sub> for annual dataset is showed. Subsequently, the summary of UNMIX, PMF and CMB sources, and average contribution to PM<sub>2.5</sub> for warm and cold season is showed as well.



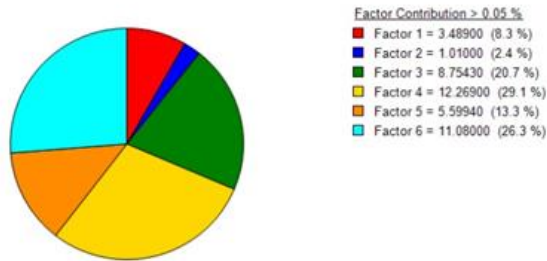
Figure 27: Summary of UNMIX and PMF sources, and average contribution to PM<sub>2.5</sub> for annual dataset (µg/m<sup>3</sup>)

### UNMIX model

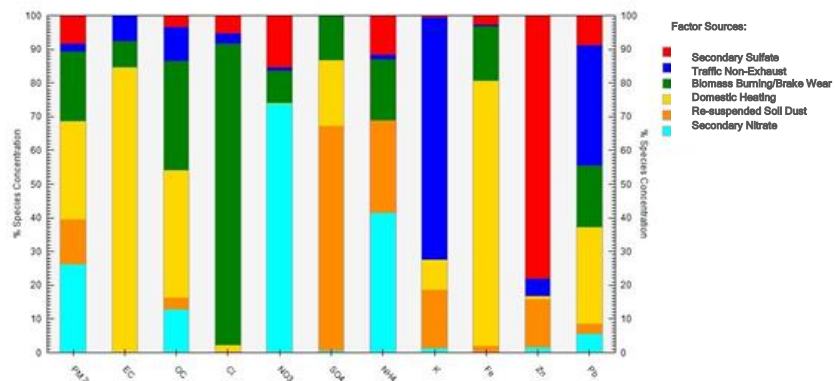


### PMF model

Distribution of sources for PM<sub>2.5</sub>

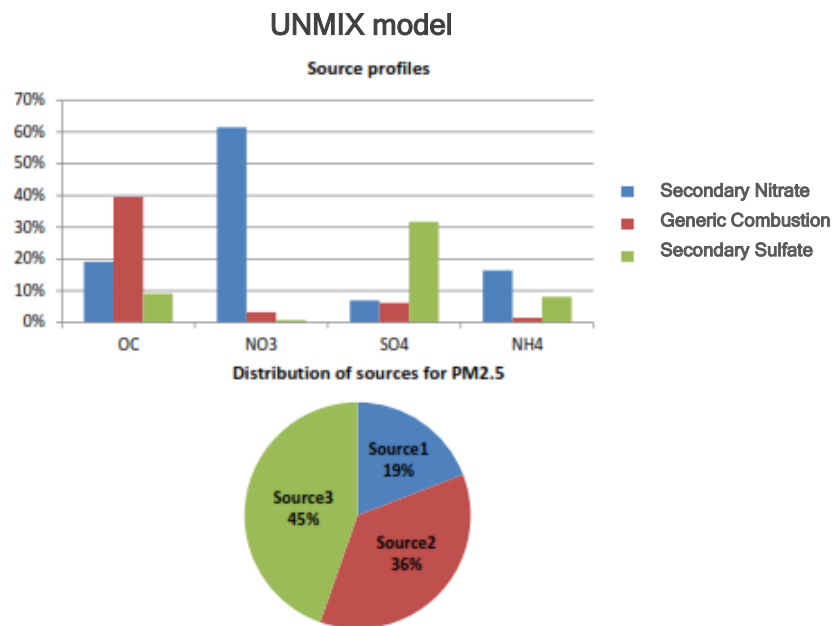


Source Profiles Fingerprint



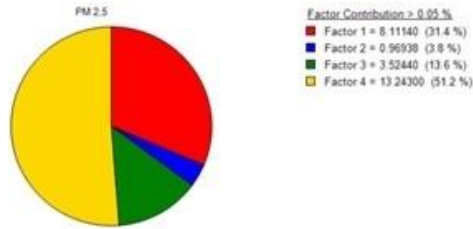
With respect to seasonal subsets, for the warm season particularly, all three models have highlighted secondary nitrate and secondary sulfate. Since, again, the number of identified source profiles is not identical; the comments about their shares cannot be drawn. UNMIX model and CMB have identified 3 source profiles, thus the only difference in their output is for the third source profile. It should be noted, that CMB model did not assign profiles for all dataset, thus approximately 18% of dataset stood unassigned. UNMIX has identified it as a generic combustion, and CMB as a traffic exhaust. The only notable difference was in the sulfate contribution. In UNMIX model the sulfate source profile has the biggest contribution towards the total PM<sub>2.5</sub> concentration mass, nevertheless, CMB as the biggest contributor identified traffic exhaust, followed by sulfate. On the other hand, PMF model has recognized 4 sources, among which traffic exhaust was identified, as in CMB model as well. However, instead of generic combustion that was the source profile identified by UNMIX, PMF recognized biomass combustion/break wear. Once again, no major differences were identified in the model output of three models, except the contributions of the secondary sulfate source. Still, alignment in the number of source profiles identified is needed for more accurate comparison.

Figure 28: Summary of UNMIX, PMF and CMB sources, and average contribution to PM<sub>2.5</sub> for warm season dataset (µg/m<sup>3</sup>)

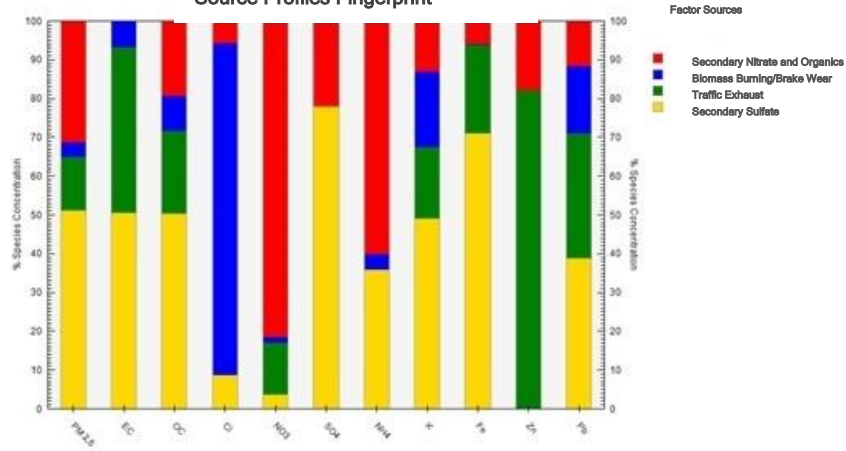


## PMF model

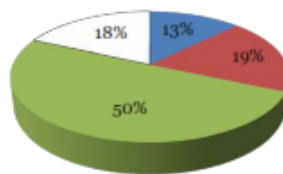
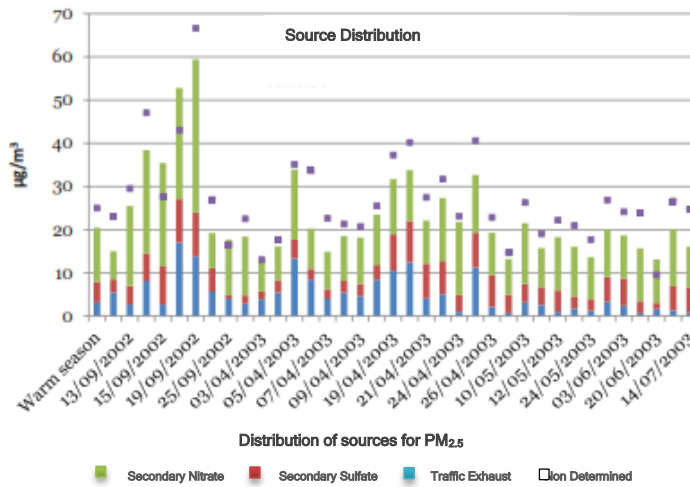
Distribution of Sources for PM<sub>2.5</sub>



Source Profiles Fingerprint



## CMB model

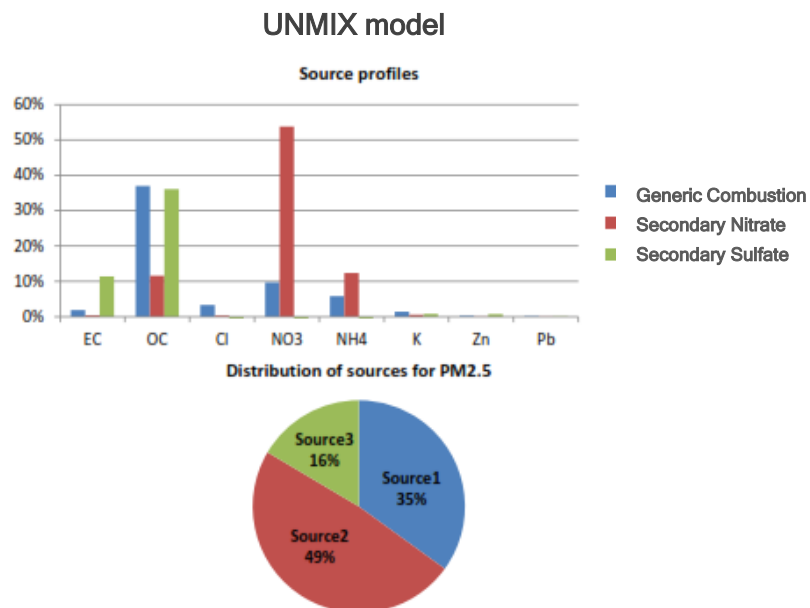


Lastly, the comparison on the winter subset was performed. In this case, UNMIX has recognized 3 source profiles, while PMF and CMB, four. The only shared source profile

among all three models is secondary nitrate. It is the biggest contributor to the total PM<sub>2.5</sub> concentration mass in both UNMIX and PMF model, while in the CBM model it is on the second place, yet very close to the first contributor. Firstly, UNMIX model did not consider sulfate in the analysis, thus secondary sulfate was impossible to be determined as a source profile; accordingly the remaining two sources were assumed to be generic combustion. As it comes to the CMB and PMF models, besides secondary nitrate, the rest of the profiles are mainly differing. In particular, CMB model recognized traffic exhaust as the first contributor, while the remaining two profiles are secondary ammonium and secondary sulfate. It should be highlighted, that CMB model did not assign profiles for all dataset, approximately 15% of dataset stood undetermined. On the other hand, PMF model recognized as the second biggest contributor domestic heating and re-suspended soil dust, and lastly, crustal matter as the minor contributor to the total PM<sub>2.5</sub> concentration mass. It may be concluded that traffic source identified by CMB is shared between re-suspended soil dust and crustal matter in the PMF model, while the secondary sulfate could be hidden in the domestic heating source.

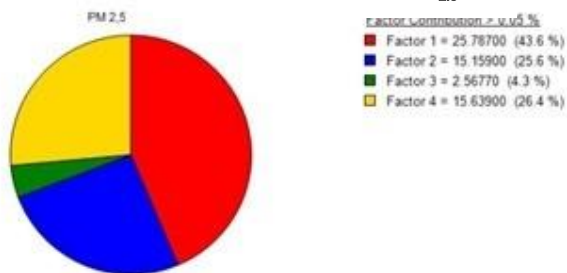
According to performed comparison, it may be concluded that all three source apportionment models have generated similar contribution for a given components; thus this observation indicates a statistically strong and reliable quantitation of the identified sources.

Figure 29: Summary of UNMIX, PMF and CMB sources, and average contribution to PM<sub>2.5</sub> for cold season dataset (µg/m<sup>3</sup>)

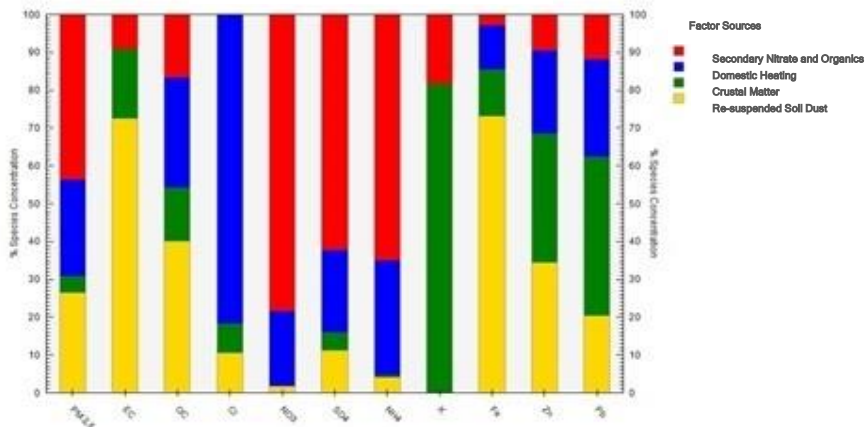


# PMF model

Distribution of Sources for PM<sub>2.5</sub>

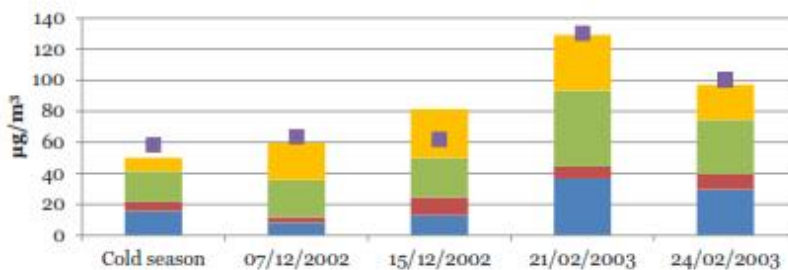


Source Profiles Fingerprint



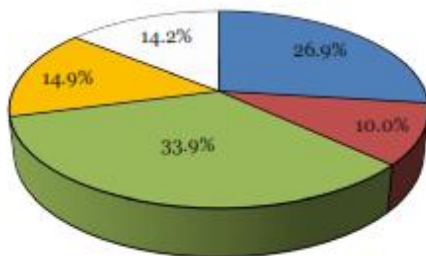
# CMB model

Source Distribution



Distribution of sources for PM<sub>2.5</sub>

Secondary Nitrate Secondary Sulfate Traffic Exhaust Secondary Ammonium Non Determined



## Conclusion

Particulate matter as being one of the air pollutants is on its rising trend of growing attention. Why? Particulate matter is well-known to its great potential for global warming and harmful effects on human health and the environment. What is worsening the situation is that these effects do not occur on the local scale only, but the global as well. World leading organizations for health are trying to raise awareness that PM affects human health more than any other pollutant, mainly towards the respiratory and cardiovascular system. Accordingly, it is noted that recently vast amount of scientific literature is investigating epidemiological, clinical and toxicological subjects. These study cases identified an immense scale of adverse health effects, starting from significant respiratory problems up to the cardiopulmonary adverse diseases and lung cancer. The severity of its impact is connected to the concentration levels in the atmospheric air, chemical composition, particle size distribution, duration of the exposure, and individual sensitivity. These studies draw a conclusion that these effects are caused by both, short-term and long-term exposure to particulate matter usually inhaled within the urban areas throughout the world. World Health Organisation states that particulate matter and effects that it causes, triggers death of about 0.85 million people annually, mainly in the developing countries.

On the other hand, particulate matter's influence on the environment it rather obvious. Fossil combustion creates fumes and vapours that are reducing visibility, while besides that, due to its chemical and physical characteristics, it influences the alteration of the radiative balance of the atmosphere's absorbing and reflecting solar radiation. Albedo, known as the reflectivity of the Earth's surface is reduced as well, mainly because of the global warming phenomenon, i.e. reduction of snow and ice (surfaces that have the most potent reflective power). Particulate matter that is suspended in the air creates fog, mainly in the urban areas, but contributes vastly to the cloud's creation as well. Clouds are formed when condensation nuclei is created, around which the water droplet forms. When the rain is created from these clouds, it may effect in forming strong fog or acid rain that will have a severe effects of corrosion on the flora and fauna, buildings, monuments and damage electrical installations. Additionally, the concentration of these particles may influence greatly on the rainfall regime. On the other hand, earth's climate is immensely influenced by the type and concentration of particulate matter. Firstly, its presence is driving global warming on a wide range, where some of the particles have enormously large global warming potential. However, that is not the only impact on the earth's climate. As mentioned before, it is very important to control levels of the Earth's reflectivity. Particulate matter as reflective substance is helping to decrease temperature on the earth with its shielding effects on the sunlight, but it keeps reflected infrared radiation below its layer as well.

Severity of the impacts that air pollution has, became more than apparent when report from European Environmental Agency was released. It claimed that dirty air cost society 189 billion euros a year; thus campaigners have demanded a crackdown on Europe's worst polluters. Referring to data from the EEA report, half of the damage is being done by only one percent of industrial plants. Coal-fired power plants and the air pollution that they emit, cost Europe approximately 189 billion euros in 2012., According to EEA figures, among all European countries, Germany was Europe's biggest contributor, It has generated 38.2 billion euros of damage to society as a result of air pollution. Subsequent country was the United Kingdom with 25.9 billion euros, followed by Poland (24.8 billion euros), France (14.3 billion euros) and Italy (13.8 billion euros). As expected, from the 30 most damaging facilities in Europe, eight were in Germany, six in Poland, four in Romania, three in Bulgaria and the United Kingdom, two in Greece, and one in the Czech Republic, Estonia, Italy and Slovakia.

Decision makers, in order to create effective strategies to control of reduce impacts on the air quality, mainly use findings of the contributions of the various emission sources to the particulate matter concentration levels. Scientific technique that examines this phenomenon is called source apportionment. This technique is having two approaches, receptor-oriented and source-oriented. Unlike other air model, receptor model do not use pollutant emissions, meteorological data and chemical transformation mechanism. Alternatively, receptor model uses chemical and physical characteristics of gases and particles measured at source and receptor to identify both, presence and contribution of sources to receptor concentrations. Since, the dataset used for this case study contained concentrations levels and chemical specification of the particulate matter that was sampled at a receptor (site) location, the used model was receptor-oriented. The main objective of the model is to identify and quantify the sources of air pollutants at a receptor location, their profiles and their relative contribution to the total particulate matter concentrations.

Receptor models were developed in the United States and are vastly used to solve air quality problems. Accordingly, the EPA<sup>7</sup> has developed several receptor models in order to reinforce the air quality management. The ones currently available are Chemical Mass Balance (CMB), UNMIX and Positive Matrix Factorisation (PMF). This case study was examined by the PMF receptor model. PMF represents a multivariate factor analysis program that decomposes a matrix of identified data set into two matrices, factor contributions and factor profiles. These factor profiles needs to be studied by the user in order to allocate the source types that may be influencing to the same cluster, by using quantified source profile information and emissions of discharge registers.

The fundamental objective of this work is to give contribution to better understanding of the nature of particulate matter and to quantify relative significance of diverse emission

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<sup>7</sup> EPA - US Environmental Protective Agency

sources and enable decision makers to convey efficient air quality remediation plans. Thus, the scope of this work is to perform a particulate matter source apportionment by means of the PMF model on the particulate matter data set from Milan. The main focus is on PM<sub>2.5</sub>, since recent air quality standards for particulate matter inquired by European Union, studies also PM<sub>2.5</sub>, besides PM<sub>10</sub>, where were specified both, concentration limits and an exposure reduction target. The second reason for centralizing on PM<sub>2.5</sub> is that anthropogenic carbon forms are almost completely included in fine particulate matter.

The original dataset have been reduced in order to enable run of the model, following the special criteria of the PMF model. Firstly the number of species was reduced from 21 up to 11, excluding mainly the trace elements. However, since some species were still missing measurement entries, exclusion of the sampling dates needed to take place as well. Thus, from starting number of 21 species and 162 concentration samples, the dataset ran in the model contained 11 species and 99 concentration samples. The investigation was firstly made at the whole dataset. Afterwards the data was separated into warm and cold season subsets, where additional investigation was made.

For what concerns the annual dataset of PM<sub>2.5</sub> in Milan, the simulation by PMF model generated 6 sources (secondary sulfate, traffic non-exhaust, biomass burning/brake wear, domestic heating, re-suspended soil dust, secondary nitrate) with average contribution to the total PM<sub>2.5</sub> mass that ranges between 10% and 20%, and as expected, significant daily variability is identified. In particular, the main contributor is identified as the domestic heating (29.1%). The one that follows is a secondary nitrate (26.3%). These two findings are more than likely to be found, since the sampling area represents very populated urban area, where great amounts of fossil fuels are being combusted. The remaining sources are biomass burning/brake wear (20.7%), re-suspended soil dust (13.3%), secondary sulfate (8.3%) and lastly, traffic non-exhaust (2.4%) source. On the plot of time series, it was identified that domestic heating contributes on a smaller scale in the warm season. Similar trend is identified for the burning of biomass/brake wear and secondary sulfate, which contribute vastly in the cold season as heating is immensely used. Re-suspended soil dust shows the biggest contribution in between the cold and warm season, while traffic non-exhaust and secondary nitrate, show the similar, relatively uniform trend, with picks in September. The sources were identified by comparing obtained relative contributions with results acquired from previous studies.

On the seasonal basis, the number of sources identified for both seasons were four. However, the type of sources identified, vary from season to season. The warm season highlighted, secondary nitrate and organics, biomass burning/brake wear, traffic exhaust and lastly secondary sulfate. On the contrary, the cold season obtained slightly different results: secondary nitrates and organics, domestic heating, crustal matter and traffic



exhaust. Domestic heating draws immediate attention as it is appearing only in the cold season, from apparent reason of the cold weather. On the other hand, traffic exhaust is present in both seasons, where it greatly contributes to the total PM<sub>2.5</sub>. That trend should draw attention of the decision makers to convey strict control and try to reduce the strong impact from this source. For the concern of the average contribution to the PM<sub>2.5</sub> mass concentration, two sources are contributing the most in the warm season: secondary sulfate (51.2%) and secondary nitrate and organics (31.4%). The ones that follow are contributing on the much smaller scale, traffic exhaust (13.6%) and biomass burning/brake wear (3.8%). As for the cold season, out of four sources, three are contributing with the great shares: secondary nitrates and organics (43.6), traffic exhaust (26.4%) and lastly domestic heating (25.6%). The smallest contribution is identified in the crustal matter (4.3%).

Due to the model structure, it is not possible to make more detailed distinction on the type of combustion. Still, some important sources were recognized and this information may be used to raise awareness as well to create some strategies with an aim to improve the air quality of the area. Looking more into results obtained, it may be concluded that the traffic exhaust contributes as double as in the warm season, growth that may be associated with reduced traffic during summer months. By investigating cold seasons, it can be remarked that no secondary sulfate source has been identified. The reason for it may be in a fact that lower temperatures do not favour formation of the sulfate. On the other hand, as expected, the combustion contribution is much higher in the cold than in the warm season, mainly due to the domestic heating emissions. Lastly, there is a slight change in contribution of secondary nitrates and organics from warm season to cold one. Nitrate is having a higher volatility at higher temperatures, thus its contribution is reduced in the warm season.



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## Annex I – How PMF Works

In receptor models data set is represented as a data matrix  $X$  of  $i$  by  $j$  dimensions, in which  $i$  identifies the number of samples and  $j$  the number of chemical species which were measured, with uncertainties  $u$ . The objective of the model is to solve the chemical mass balance (CMB) between observed concentrations and source profiles, as presented in Equation; with number of factors  $p$ , the species profile  $f$  of each source, and the quantity of mass  $g$  that is created by aggregation of each factor to each individual sample.

$$x_{ij} = \sum_{k=1}^p g_{ik} f_{kj} + e_{ij} \quad (1)$$

Where  $e_{ij}$  is the residual of sample or species. The CMB equation can be resolved using several models, among which are EPA CMB, EPA Unmix and EPA PMF.

Representing the multivariate factor analysis method, PMF divides original data set into two matrices: factor contributions (G) and factor profiles (F). Analyst is then processing these factor profiles in order to identify the source types that may be contributing to the sample by using information derived from the measured source profile and emission or discharge inventories.

PMF uses both user-provided uncertainties associated with the sample data and as well sample concentration in order to measure weight single points. By having this characteristic, analyst is in power to account for the confidence in the measurement. For instance, data under the detection can be preserved for use in the method, with the assigned uncertainty adjusted so that these individual points from data set have less impact on the result than measurements above the detection limit.

Using the PMF model minimization of the objective function  $Q$ , factor influences and profiles are derived (Equation2).

$$Q = \sum_{i=1}^n \sum_{j=1}^m \left[ \frac{x_{ij} - \sum_{k=1}^p g_{ik} f_{kj}}{u_{ij}} \right]^2 \quad (2)$$

$Q$  represents a critical factor for PMF and during the model runs, two version of  $Q$  are showed:

- $Q(\text{true})$  indicates the accuracy of factor calculated including all points
- $Q(\text{robust})$  indicated, on the other hand, the accuracy of factor calculated excluding points not fit by the model, and identified as samples for which the uncertainty-scaled residual shows values greater than 4.



The differentiation between  $Q(\text{true})$  and  $Q(\text{robust})$  create a measure of the impact of data points which have high scaled residuals. Usually peak influences from sources that are not present during the sampling time period are connected with these points. However, in some cases uncertainties may be too high which implies similar  $Q(\text{true})$  and  $Q(\text{robust})$  values because the residuals are scaled by the uncertainty.

PMF model requires various iterations of the Multilinear Engine (ME) in order to support identification of the most optimal factor influences and profiles. This is needed because of the nature of the ME procedure that initiates the search for the consistent factor profiles with a randomly created factor profile. Additionally, this factor profile is adjusted by the gradient approach with an aim to graph the ideal path to the best solution. In other words, the model creates a multidimensional space using the measurement samples and then passes through the space using the gradient approach to arrive to its final destination of the best result along the path. Identification of the best solution usually implies the lowest  $Q(\text{robust})$  value along the path (i.e. the minimum  $Q$ ) and it may be perceived as the bottom point of the through in the multidimensional space. Since the random starting point is determined by the seed valued and the path which dictates it, there is no assurance that the gradient approach will always identify the lowest point in the multidimensional space (global minimum), but instead, it may find the local minimum. In order to increase the chance of reaching the global minimum, the model should be run at least 20 times for the development of the solution and 100 times for the final solution, with a choice of different starting point every time.

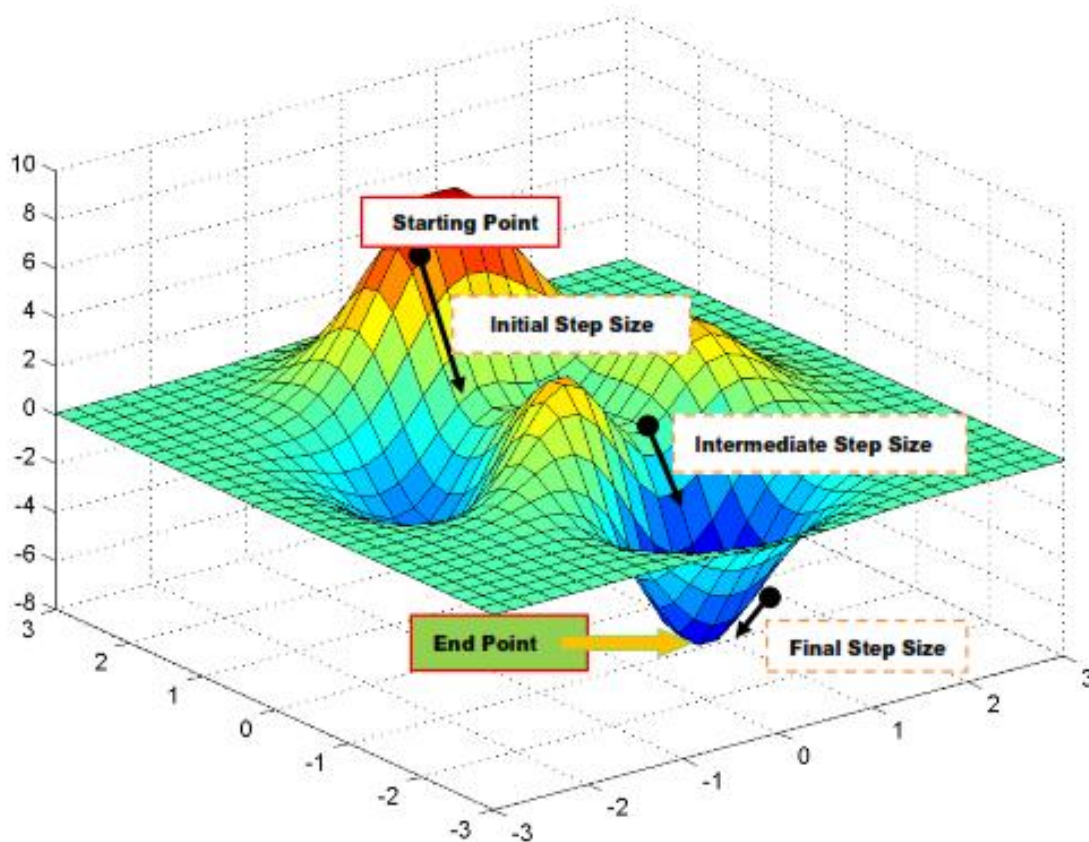
The points that are not fit by the PMF does not influence the  $Q(\text{robust})$ , it is normally used as a parameter for a choice of the optimal run from the various ones. Likewise, the variability of  $Q(\text{robust})$  gives an information if the starting run solutions have great variability, because of the random seed which is used to initiate the gradient algorithm in diverse positions. If the provision of the stabilized path towards the minimum is obtained by data used,  $Q(\text{robust})$  will have only little variations between the runs. On the other hand, the arrangement of the starting point and the space defined by the data set will impact the direction to the minimum value, resulting in that way in diverse values of  $Q(\text{robust})$ . The best optimal solution is created by the lowest value of the  $Q(\text{robust})$ , and it is used by default.

Another influence in factor profiles among PMF runs can be noted if some chemical transformations or changes in processes appear. In order to evaluate the variations between runs, two approaches are available: intra-run residual analysis and a factor summary of the species distribution compared to those of the lowest  $Q(\text{robust})$  run. Evaluation of all the error estimates in PMF is performed by the analyst. Lastly, the variation in the PMF results can be assessed with the following three methodologies:

- Bootstrap (BS) analysis is used to identify if there exist a small set of observations that can disproportionately impact the result. BS error intervals which includes effects from random errors and as well effects of rotational ambiguity. The presence of infinite results that are very much alike to the result created by PMF is causing rotational ambiguity. In other words, for any present pair of matrices there can be infinite variations of the pair that are created by a single rotation. Rotation of the space is impossible to be limited, since only boundary condition of non-negative source contribution is present. These errors are not affected by the user's sample uncertainties and are normally representing robust errors.
- Displacement (DISP) analysis helps the analyst to comprehend the final result in more detailed aspect, including its sensitivity to minor variations. Error intervals from this analysis method include effects of rotational ambiguity, however does not include impacts of random errors in the data set. Drawback of this analysis is that data uncertainties can directly influence DISP error estimates. Thus, ranges for down-weighted species are most likely to be large.
- BS-DISP (hybrid approach) error intervals comprise influences of random errors and rotational ambiguity. BS-DISP solutions are more robust than DISP solution, because the DISP phase of BS-DISP does not relocate as great as DISP by itself.

Multi-linear Engine (ME) resolves a range of multi-linear PMF problems comprising bilinear, trilinear, and mixed models. It is created to solve relevant problems by joining two steps. The analyst creates a table that outlines the PMF model itself. This step is followed with automated secondary program that interprets the tabulated model parameters and calculates the results. By using EPA PMF model for the solution of the PMF problem, the primary step is resolved with an input file that is created with the help of EPA PMF user interface. When model has been defined, data and user details are inserted into the secondary ME program by EPA PMF. This program resolves the PMF program iteratively, minimizing in the same time the sum-of-squares object function, Q during the various series as indicated in the Figure11. When diminishing solutions for minimizing Q are obtained by processing additional iterations, it means that a stable solution has been reached. Filtering of the possible solution produces a finer scale over three stages of iterations.

Figure 30: Conjugate Gradient Method



The first stage of iteration assigns the general range of solution in space. For the good results, the change in  $Q$  ( $dQ$ ) is required to be less than 0.1 over 20 consecutive steps in less than 800 steps. Subsequent step identifies the neighborhood of the final solution. In this case,  $dQ$  should be less than 0.005 over 50 consecutive steps in less than 2000 total steps. The final step converges toward the best possible  $Q$ -values, following as well that  $dQ$  should be less than 0.0003 over 100 consecutive steps in less than 5000 steps.

When small data sets are being used, i.e. less than 300 observations, generally a few hundred iterations is needed. However, for bigger data sets around 2000 iterations are needed in order to achieve accurate solutions. In a case that a solution is found but do not meet any of the requirements set by three stages, then a solution is non-convergent.

Outcome of the ME process is being interpreted by PMF and then modified for easier user interpretation. The 5.0 version of the EPA PMF uses the newest ME and a PMF script file. These two approaches were developed by Pentti Paatero at the University of Helsinki and Shelly Eberly at Geomatic Tools.

The PMF model requires a data set that contains a range of parameters which have been measured across various samples. More specifically, PM<sub>2.5</sub> data set comprises of 10 to 20 species over 100 samples. However, an uncertainty data set is also necessary for the complete calculation. This data set is calculated by using propagated uncertainties or some similar accessible information, for instance collated sampling precision.

Application of the PMF is connected to a wide range of data, including 24-hour PM<sub>2.5</sub>, size resolved aerosol, deposition, air toxics, high time resolution measurements, i.e. aerosol mass spectrometers (AMS) and volatile organic compound (VOC) data.